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Graphical abstract

Cyclic variations of sulfate and boron concentrations and isotopes in deep groundwaters in the Aquitaine Basin, France

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Abstract

Concentrations and isotope contents of major and trace elements are principal factors in determining the origin of the chemical composition of groundwater. This paper focuses specifically on the use of sulfur and boron isotopes to characterize the origin of cyclic variations in the deep Eocene aquifer in the Aquitaine sedimentary basin (southwest of France). It is part of a multi-layer system mainly composed of sands and sandstone deposits. Groundwater contained in this deep reservoir is known to present stable chemical compositions, allowing its use for various purposes like drinking water, geothermal energy, thermal activity and agricultural irrigation. However, among the dozens of wells exploiting this aquifer and despite the reservoir's substantial depth, variations in sulfate concentration have been identified in a limited area of the reservoir. These fluctuations are cyclic and they seem to be correlated with water level variations due to gas storage activities nearby in the same aquifer. Regular water samplings and analyses of major and trace elements and their isotopes have identified that bore concentration variations are correlated with sulfate variations. A geochemical modelling approach based on water mixes elucidates the causes of these variations in chemical composition, especially the boron and sulfate concentrations and their respective $\delta^{11}\text{B}$ and $\delta^{34}\text{S}_{\text{SO}_4}$ values. From these numerical results, we identify that different sources explain the variations of boron concentration, one part coming from the silicates alteration (present in sub-layers of the exploited aquifer) and the other part coming from the evaporites alteration (present in the underlying molasse unit). These results also confirm the existence of mass and potentially water transfers between the different sub-layers of the reservoir and with the underlying molasse aquitard, implying new constraints for the future hydrogeologic modelling.

Keywords: deep aquifer; France; groundwater; sulfur stable isotopes; boron stable isotopes; mixing geochemical models.

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1 - Introduction

The Aquitaine basin is a wide sedimentary basin located in the southwest of France. It is constituted of aquitards and aquifers used for different purposes. Because of the various uses of the aquifers, the aquifer needs to be efficiently managed and the resources have to be investigated. To achieve this, public and private institutions initiated recent investigations on the Eocene aquifer through a scientific project named GAIA (Wuilleumier et al., 2018). This study embraces numerous areas of the earth sciences, including geology (Ortega et al., 2017), hydrogeology and hydrogeochemistry (Wuilleumier et al., 2018; Gal et al., 2018 a,b; André et al., 2018).

For the hydrogeochemistry research, the main goal is to investigate in detail the origin of water mineralization. The water chemistry in the Eocene aquifer (Aquitaine Basin – Southwest of France - Figure 1) has been investigated through different studies. Blavoux et al. (1993) were the first to investigate the isotope composition of the Eocene reservoir waters with characterization of ^2H , ^{18}O , tritium, noble gases and carbon-14 activity. They defined that waters have a meteoric origin with a late Pleistocene origin. The authors concluded their article that the reservoir is a complex system and they did not exclude a recharge through the overlying molassic unit. They also supposed that additional investigations on sulphur components (including isotopes) could bring information on the origins for water mineralization. In agreement with these conclusions, isotopic analyses have helped to identify geochemical basins whose properties modify the chemical water composition (André, 2002; André et al., 2002; Douez, 2007). These studies showed that the $\delta^{34}\text{S}$ of gypsum sampled in the molasse unit was consistent with $\delta^{34}\text{S}$ values of dissolved sulfate in the reservoir waters. The study confirms the hypothesis of Blavoux et al. (1993) on the potential transfers between the molassic aquitard and the reservoir. Hydrogeochemical investigations continued with the CARISMEAU project (Négre et al., 2007; 2009). During this study, the $\delta^{34}\text{S}$ of sulfate was also investigated in the Northern part of the reservoir (Entre-deux-mers area), coupled with other parameters like $^{87}\text{Sr}/^{86}\text{Sr}$ ratio (Brenot et al., 2015) or with fluoride measurements (Malcuit et al., 2014) to investigate the mineralization origin of waters. However, despite these investigations, some mechanisms are still unknown. Indeed, all the previous chemical investigations showed that the chemical composition of waters from this deep aquifer (about 500–1000 m depth) does not change usually with time. However, seasonal variations in sulfate concentrations occur in a specific area of the aquifer (Nogaro 2 well – Figure 2). About 200 samples collected at this well over more than 20 years show variations between 10 and about 100 $\text{mg SO}_4\cdot\text{l}^{-1}$ (sometimes higher). These varied sulfate concentrations are correlated with the water conductivity and anti-correlated with

89 piezometric variations (reaching about 80 m per year), which are due to the impact of gas
90 storage activities occurring within the same aquifer.

91 During this study, additional hydrogeochemical monitoring was implemented for two years to
92 increase our understanding of the mechanisms underlying these sulfate fluctuations. We
93 focused on how sulfate concentration changed over time but also investigated the other major
94 elements, traces and isotopes in the waters of the Nogaro 2 well, which exploits the Eocene
95 aquifer. The Nogaro 2 well (reference: BSS002EFSB - see Infoterre, 2020a) is used to supply
96 drinking water to the city of Nogaro (Gers Department). It was drilled in 1982 and has been
97 monitored regularly for chemical composition of groundwater. Two other wells located
98 respectively 12 and 15 km from the Nogaro 2 well, were also monitored over the same period in
99 order to compare the potential seasonal variations in the waters' chemical composition. Also
100 used for drinking water supply, the Eauze well (reference: BSS002EGAG - see Infoterre, 2020b)
101 and the Dému well (reference: BSS002EFZJ - see Infoterre, 2020c) pump within the same
102 aquifer and are also influenced by the hydrogeologic conditions of this zone.

103 Chemical analyses focused on major and trace elements but also stable and radioactive
104 isotopes. In coherence with former analyses, the new investigations confirm the sulfate
105 variations (both total sulfur and ^{34}S concentration). Isotopes of dissolved sulfate are widely used
106 in large basins worldwide for identifying sulfur origins in water (De Caritat et al., 2005; Li et al.,
107 2011; Einsiedl et al., 2015). ^{34}S in sulfate is a geochemical tracer for obtaining information on
108 the origin of sulfur and biological processes that may alter sulfate compositions in water (such
109 as bio-sulfato-reduction). As part of this study, this isotope was tracked over time to better
110 understand the origin of sulfate variations (André et al., 2002; Négrel et al., 2009). The new
111 analyses on Nogaro 2 has also revealed remarkable boron variations (both total boron and ^{11}B
112 content) correlated with sulfur variations. ^{11}B is a tracer used to define the origin of boron in
113 solution, mainly the characterization of water/rock interactions in sedimentary basins (Négrel et
114 al., 2012; Deiana et al., 2020), weathering processes in surface and subsurface systems
115 (Lemarchand and Gaillardet, 2006; Ercolani et al., 2019), pollution sources and/or the anthropic
116 inputs (Harkness et al., 2018) and salinization of coastal aquifers (Cary et al., 2015). The $\delta^{11}\text{B}$
117 value is also impacted by adsorption processes with a preferred adsorption of the light isotope
118 (^{10}B) and an enrichment of the solution in heavy isotope ^{11}B (Palmer et al., 1987; Godfrey and
119 Alvarez-Almado, 2020). At last, carbon-13 is a geochemical marker of water-rock interactions
120 (including carbonates) and carbon-14 activity is used as an indicator of the apparent age of
121 water.

Our goal in this paper was to document the mechanisms controlling the variations in chemical composition in the groundwater in this specific area and to identify potential causes of these cyclic disturbances. After interpreting the chemical analyses, we carried out geochemical modelling to characterize, from a mixing model, the different end-members (waters of different horizons) that could be responsible for these concentration variations.

2 – Geology, mineralogy and hydrogeological settings

2.1 - Geology

The Aquitaine Basin is a large sedimentary basin limited in the east by the foothills of Montagne Noire, in the south by the North Pyrenean Piedmont, in the west by the Atlantic Ocean and in the north by the Poitou Plateau (Figure 1).

The area is characterized by a thick layer of detritic deposits (molasse) aging from late Eocene to Miocene (Figure 2) and originating from the surrounding landforms (Massif Central, the Montagne Noire and the Pyrenees mountains). This deposit starts at the beginning of the Priabonian, when a drop in sea level led to a progression of late orogenic continental molassic sediments (Orthiz, 2019). The development of the Campagne formation, mainly marly, corresponds to a general regression and thus mark the continentalization of most of the Adour basin, a sub-basin located in the southwest of Aquitaine Basin (Serrano, 2001). The lake and river facies of these continental deposits have many evaporitic levels and numerous gypsum indices were found in the Campagne formation (Priabonien).

Figure 1 – Location map. The area of interest corresponds to the region investigated in the GAIA Project.

This formation overlies (Douez, 2007) the so-called Infra-Molassic Sands (SIM) of the early to mid-Eocene (Figure 2), which is mostly consisting of the Lussagnet sands of the Cuisian to Lutetian age, a prograding deltaic formation, and the underlying Nummulites sandstone of the Ypresien age, integrated into the Baliros sands formation. That sandy Eocene formation lies generally upon a layer of marl and clay deposits of the early Eocene and late Paleocene, which thickness usually ranges from 100 to some hundreds of meters (Beicip, 1984; Angrand, 2017), but, according to the location in the basin, it can lie directly upon the limestone and dolomitic limestone of the Paleocene (Douez, 2007).

Figure 2 – Geological map (to 1 million) of the area and location of the three sampled wells: Nogaro 2, Eauze and Dému. Black line NS is a geological South-North schematic cross-section (adapted from Douez, 2007).

The depth of the top of the Eocene sand layer shows notable variations according to the location in the basin. The Nogaro 2 well corresponds to a low point with a depth of about 920 m. The depth of the sand layer varies for many reasons. First, the thicknesses of the deposits may differ from one point to another in the sector depending on the amount of sediment stored during the progradation of the fan delta westward during the Middle Eocene, i.e. during the maximum peak of the Pyrenean compression, which corresponded to the maximum sediment discharge. In this type of context, the rate of movement of the delta position and the amount of detritus can change. Second, a massive but heterogeneous subsidence impacted the entire basin during the Pyrenean orogeny. During the Pyrenean compression responsible for this subsidence, the propagation of tectonic stresses also generated significant deformations in the South Aquitaine basin. At the origin of a “piercing” diapirism (particularly in the western part of the basin: Dax, Tercis, Bastenne-Gaujac, etc.) or of a “blind” diapirism (like in Audignon, Lacq, Meillon), these long wavelength deformations affect part of the tertiary deposits, in particular by installing anticline structures (Bourrouilh et al., 1995; Serrano, 2001; AGSO and BRGM, 2018).

2.2 - Hydrogeology

The Eocene aquifer is a major aquifer in the Aquitaine sedimentary basin and it is used for uses such as drinking water, field irrigation, thermalism and spas, gas storage and as a thermal resource. This aquifer extends over 150 km from east to west and 200 km from south to north and constitutes a part of a multi-layer system. The Eocene sands aquifer is covered by several hundreds of meters of Tertiary molasse unit of low permeability. Therefore, the Eocene aquifer is confined over the studied area. Part of its outcrops are located in the South, close to the Pyrenees Mountains, and in the East, in the Montagne Noire region (Figure 1). At least, the mean annual air temperature at Mont-de-Marsan is 13.6°C for the 1981-2010 period (MeteoFrance, 2020).

The average thickness of the Eocene aquifer varies according to the position in the basin from about 50 m to more than 200 m. At the Nogaro 2 well, the thickness is close to 180 m. The lithostratigraphic cuts obtained in the Nogaro 2 well clearly identify the presence of the Lussagnet sands aquifer (between 920 and 1001 m deep), which, from a hydrogeological point of view, forms the upper part of the Eocene sand aquifer in that area. The Lussagnet sands aquifer overlies the Baliros sands aquifer, which is the productive zone of the reservoir (between 1001 and 1092 m depth). This last part of the aquifer lies above a bioclastic calcareous layer (Horbaziou Formation) of low productivity (Infoterre, 2020a).

At Eauze, the Eocene aquifer is identified between 471 and 573 m. The upper part is constituted of clay interlayers and the real productive part of the reservoir (i.e. the exploited portion) lies between 504 and 573 m. It is made of sands and gravels of varying coarseness and overlies a grey calcareous horizon constituted of marls and sandstones with calcareous cement. At Dému, the well taps the Eocene aquifer between 740 and 780 m. The reservoir is overlayed by 12 m of sands and gravels and Nummulites sandstone seems to lie beneath, but the well was not drilled below this depth and the nature of the lower horizons is unknown.

The porosity of the quartz sand deposit is estimated at 20–35% (Housse and Maget, 1977; Douez et al., 2006). The average permeability (estimated from aquifer testing and modelling results) depends on the location and is generally accepted to range between $1.5 \cdot 10^{-4}$ and approximately $3 \cdot 10^{-5} \text{ m.s}^{-1}$ (Labat, 1998; Seguin, 2003). The average interstitial velocity, using a gradient of 0.001 and an effective porosity of 20%, is close to 5 m.y^{-1} . Groundwater flow is mainly oriented from SE towards NW but outflow from the aquifer is not fully known. The estimated apparent age of the groundwater (using ^{14}C data and ^{13}C data for corrections with the Pearson and Hanshaw model, 1970) is close to 25–35 ky (André et al., 2018), which is consistent with the effective advection calculated above.

The part of this aquifer that we studied is impacted hydraulically by gas storage at the Lussagnet and Izaute sites. With a maximum authorized capacity of 6.5 billion m³, the two storage facilities represent almost a quarter of France's underground storage capacity, supplying natural gas to the southwest of France and also feeding other French and European networks. The Lussagnet site has been implemented in 1957, prior to any piezometric measurements, and the Izaute site in 1981, which influence of the static water level is recorded since that date. Gas is stored on a relatively constant way throughout the year but monthly consumption in the winter can be almost five times as high as in the summer months (TEREGA, 2020). Therefore, water pressure in the aquifer is submitted to high seasonal fluctuations (more than 10 bars), inducing groundwater level variations around the storage zone. Fluctuations are recorded all around the storage sites, like in Nogaro 2, Eauze and Dému, where static water level variations can reach several tens of meters for these three wells (Figure 3).

Figure 3 – Static water level recorded at three wells. Top: Nogaro 2; Middle: Eauze; Bottom: Dému. The water level recording in Eauze starts in 2007, year of the wellbore drilling.

2.3 - Mineralogy

The mineralogy of the Eocene sands aquifer is relatively simple, containing mainly quartz augmented with calcite, and, in some places, dolomite and K-feldspars (André, 2002). In the Nogaro 2 well but also in other wells located in the zone of interest, pyrite indices were recognized sometimes in the Lussagnet sands and sometimes under the base of the Lussagnet sands aquifer, in the clay part of the Nummulites sandstones (Medium Eocene). More precise attempts to characterize this pyrite were made during this study in order to determine the $\delta^{34}\text{S}$ value of these sulfide minerals. Cuttings from a recently drilled well a few kilometres away from Nogaro 2 were used to do this characterization. XRD analyses of these cuttings revealed the presence of traces of pyrite (about 1% in mass). Extractions were then carried out in laboratories. However, it has proved impossible to isolate the pyrite properly to carry out isotopic analyses. The small amount, the grain size, and the pyrite being coated by quartz did not allow further analysis.

The mineralogy of the molasse sediments, although less well investigated, is richer, with quartz, feldspars, mica and several clay types (detrital limestone levels with sandy-clay deposits). At the lower part of the molasse deposits, crystallised gypsum has been observed in cuttings at many places.

251

252 **3 - Water chemistry**

253 This study is based on two types of data: complete measurements of water chemical
254 composition recorded between 2016 and 2018 and historical data, corresponding to sulfate
255 concentrations recorded for more than 20 years.

256 **3.1 - Sampling and analytical methods**

257 Nine sampling campaigns (April, July, October 2016, January, March, June, October, December
258 2017 and March 2018) and chemical analyses were carried out over two years. The
259 investigated parameters measured on-site were physico-chemical data: temperature,
260 conductivity (standardized to 25 °C), redox potential (Eh), pH. Parameters are measured at the
261 wellhead, after purging the borehole. Water is collected only when physico-parameters are
262 stable.

263 The water samples were collected in polyethylene bottles as raw samples for $\delta^{18}\text{O}$ - $\delta^2\text{H}$, ^{13}C
264 analyses and ^{14}C activity measurements, and filtered through 0.45 μm PVDF filters using a
265 Nalgene filter apparatus for chemical and boron, sulfate and strontium isotopes analyses.
266 Bottles dedicated to cation and strontium isotopes analysis were acidified with 15N ultrapure
267 HNO_3 to $\text{pH} < 2$. A cadmium acetate solution was added to water samples dedicated to sulfate
268 isotopes analysis to precipitate sulfides, decrease pH and eliminate potential bacteria. Samples
269 were stored at 4 °C in the dark prior to analysis.

270 Major element concentrations are measured in BRGM's laboratory: Ca, Mg, Na, K and silica
271 concentrations were measured by Inductively Coupled Plasma – Atomic Emission Spectroscopy
272 (NF EN ISO 11885 of November 2009) and Cl, SO_4 anions were analysed by ionic
273 chromatography (Standard NF EN ISO 10304 of 1 July 2009). The accuracy for all element
274 concentrations is 5 %, except Na and K with 10 %. Alkalimetry was measured by potentiometry
275 and HCl titration according to Standard NF EN ISO 9963-1 February 1996, with an accuracy of
276 15%.

277 Trace element concentrations (Al, B, Ba, Li, Mn, Sr, Br) were measured by inductively coupled
278 plasma mass spectrometry according to Standard NF EN ISO 17294-2 of April 2005, with
279 variable accuracy: Al 40%, B 15%, Ba 5%, Li 10%, Mn 10%, Sr 10% and Br 10%. Total iron
280 (Fe_{total}) concentration was measured by Inductively Coupled Plasma – Atomic Emission
281 Spectroscopy (accuracy 10 %) and F by ionic chromatography with an accuracy of 5 %.

282 Accuracy and precision for major and trace elements was verified by repeated measurements of
283 standard materials during the course of this study: namely Merck etalon (1 g.l^{-1}) for Ca, Mg, Na,

K, Fe, SiO₂, Cl, F and a multi-element solution (10 mg.l⁻¹) from Inorganic Ventures for trace elements.

Stable isotopes concentration ratios (²H/¹H and ¹⁸O/¹⁶O) of water molecules were analysed with a Finnigan® MAT 252 mass spectrometer connected to an automatic device, using the gas equilibration method (H₂ for hydrogen and CO₂ for oxygen). The isotopic ratios are expressed in ‰ vs V-SMOW (Vienna Standard Mean Ocean Water). Precisions are 0.8 ‰ for δD and 0.1 ‰ for δ¹⁸O.

For ³⁴S_{SO4} and ¹⁸O_{SO4} isotopic analysis, sulphides (precipitated as CdS after cadmium acetate adding) are first removed by filtration on 0.22 µm membrane. Then, dissolved sulfates are precipitated as BaSO₄ at pH < 4 by adding a BaCl₂ solution. The precipitate of BaSO₄ is then recovered on a 0.45 µm filter and dried at 50°C. For δ³⁴S_{SO4}, 250 µg of BaSO₄ are mixed with V₂O₅ in a tin capsule, and injected in a flash combustion elemental analyzer (Flash EA) where BaSO₄ is reduced to SO₂ at 1700°C. The SO₂ gas, purified by gas chromatography is then analyzed by CF-IRMS (Continuous Flow Isotope Ratio Mass Spectrometer) on a Thermo Scientific Delta V Plus spectrometer. For δ¹⁸O_{SO4}, 150 µg de BaSO₄ are placed in a silver capsule and injected into a graphite pyrolysis oven in a high temperature conversion elemental analyzer (TC/EA) at 1450°C. The CO gas resulting from the reaction between oxygen and graphite is purified, and analyzed by CF-IRMS on a Thermo Finnigan Delta Plus XP spectrometer. Calibration ranges (from -34.05 ‰ to +21.12 ‰ for δ³⁴S_{SO4}, and from -11.35 ‰ to +12.13 ‰ for δ¹⁸O_{SO4}) are realised with three international standards (IAEA SO5, IAEA SO6, and NBS 127). The S and O isotope compositions are reported in the usual δ-scale in ‰ with reference to V-CDT (Vienna Canyon Diablo Troilite) and V-SMOW (Vienna Standard Mean Ocean Water). The sulfate-isotope compositions (δ³⁴S_{SO4} and δ¹⁸O_{SO4}) were measured with a precision of ±0.3 ‰ vs. V-CDT for δ³⁴S_{SO4} and ±0.5 ‰ vs. V-SMOW for δ¹⁸O_{SO4}.

Boron isotopic compositions were determined on a Finnigan® MAT262 solid source mass spectrometer in a dynamic mode. For these samples, water volumes corresponding to a boron quantity of 10 µg underwent a two-step chemical purification using Amberlite IRA-743 selective resin. The boron aliquot sample (2 µg) was then loaded onto a Ta single filament with graphite, mannitol and Cs, and the B isotopes were determined by measuring the Cs₂BO₂⁺ ion. Total boron blank is less than 10 ng. Purified samples are always analysed twice. The values are given using the δ-notation (expressed in ‰) relative to the NBS951 boric acid standard. The ¹¹B/¹⁰B of replicate analysis of the NBS951 boric acid standard after oxygen correction was 4.05387 ± 0.00120 (2σ, n = 192) during this period (April-2016 to May-2018). The reproducibility of the δ¹¹B determination is then ± 0.3‰ (2σ). The internal uncertainty is often better than 0.2‰

($2\sigma_m$). Long-term accuracy and reproducibility of the overall procedure were verified by the repeated measurements of the IAEA-B1 seawater standard (Gonfiantini et al. 2003) for which the mean $\delta^{11}\text{B}$ value obtained is $39.20\text{‰} \pm 0.30$ (2σ , $n=93$) in accordance with the accepted value for seawater.

Chemical purification of Sr was performed with an ion-exchange column (Sr-Spec) before mass analysis according to a method adapted from Pin and Bassin (1992), with total blank <0.5 ng for the entire chemical procedure. After chemical separation, around 150 ng of Sr was loaded onto a tungsten filament with tantalum activator and analyzed with a Finnigan® MAT262 multi-collector mass spectrometer. The $^{87}\text{Sr}/^{86}\text{Sr}$ ratios were normalized to a $^{86}\text{Sr}/^{88}\text{Sr}$ ratio of 0.1194. An average internal precision of ± 10 ppm ($2\sigma_m$) was obtained and reproducibility of the $^{87}\text{Sr}/^{86}\text{Sr}$ ratio measurements was tested through repeated analyses of the NBS987 standard for which we obtained, during the overall duration of this study, a mean value of 0.710246 ± 0.000011 (2σ , $n = 294$). Sample ratios were normalized to the certified value of the NBS987 (0.710240).

Carbon stable isotopes ($^{13}\text{C}/^{12}\text{C}$) and ^{14}C activity were analysed by Beta Analytic Inc. Water was first acidified with orthophosphoric acid under vacuum to get CO_2 . The gas is purified and a small aliquot fraction was used for the measurement of $^{13}\text{C}/^{12}\text{C}$ ratio by Isotopic Ratio Mass Spectrometry (IRMS). Measured $^{13}\text{C}/^{12}\text{C}$ ratios are calculated relative to the PDB-1 standard. CO_2 was reduced by H_2 into graphite over a cobalt catalyst. The ^{14}C measurements were performed by Accelerator Mass Spectrometry (AMS). The analytical result (in pMC) is obtained by measuring sample $^{14}\text{C}/^{13}\text{C}$ relative to the $^{14}\text{C}/^{13}\text{C}$ in Oxalic Acid II (NIST-4990C). The typical analytical uncertainty is about 0.1 pMC (1σ).

3.2 - Historical data

Nogaro 2 is a well used to provide drinking water. Because of this use, the water quality is regularly controlled. Since the sulfate concentration of waters exceeds sometimes drinking waters regulations (in France, the breakthrough value of sulfate concentration is 250 mg.l^{-1}), a monthly monitoring is organized by the Regional Health Agency and monitoring is managed by the laboratory of the Gers department. However, this monitoring only focus on electrical conductivity and sulfate concentrations measurements without any information on other major and trace elements. All the chemical data acquired by both laboratories are registered on ADES website (ADES, 2020), a national database available for consultation. This hydrogeochemical monitoring revealed the variable sulfate concentration over time (Figure 4). From 2003-2015, between 6 and 12 measurements were made per year. While the chemical composition of the waters of the Eocene Sands aquifer varies only moderately over time, the waters show seasonal variation in sulfate concentrations. The lowest concentrations were of the order of 10

mg.l⁻¹, while the strongest reached values greater than 100 mg.l⁻¹. Although sulfate was monitored regularly, available water analyses were often incomplete and it was difficult to correlate these variations in sulfate content with variations in other chemical elements.

Figure 4 – Sulfate concentrations for waters in the Nogaro 2 well between 1980 and 2016

A visual comparison of piezometric levels and measured concentrations shows that high sulfate concentrations are found when piezometric levels are low (Figure 5). There is therefore an apparent anti-correlation between water level measurements and sulfate concentration measurements. We therefore assume the preeminent role of the pressure variations in the aquifer. We also observe that the shape of concentration cycles are different from the shape of hydraulic head variations: “concentration plateaus” of variable length separate the sulfate peaks whereas the hydraulic head looks like a sinusoidal curve. So the sulfate plateau reveals a minimum concentration of this element (10 mg.l⁻¹) in the aquifer. This 10 mg.l⁻¹ value does not correspond to analytical limitations but to a real value of sulfate concentration in the reservoir.

Figure 5 – Water levels (open black circle) and sulfate concentrations (black triangles with grey line) for water from the Nogaro 2 well between 2003 and 2013

The Eauze well has been drilled quite recently (in 2007) and there has not been regular monitoring of sulfate concentration for this well (only 4 data points are available over the period 2007-2016). However, given the piezometric amplitudes over this period and the geographical proximity to the Nogaro 2 well, the question arises of how dissolved sulfate behaves in the water from this well.

The Dému well also has high piezometric amplitudes (about 30 m - Figure 3). However, based on the available data before this study, no correlation between piezometric level and concentrations could be found.

Consequently, the chemical cyclic behaviour observed at the Nogaro 2 well and potentially in other wells of the sector raises many questions on the origin of these fluctuations. However, given the available data and particularly the small number of complete chemical analyses, we decided to carry out quarterly monitoring of the chemical composition of the waters of the three wells. The aim was to be able to understand the reason for these variations, both in terms of chemistry and as an inferred consequence of the system's hydrodynamic behaviour.

3.3 - Results

3.3.1 - Physico-chemical parameters

Wellhead measured temperatures – reaching up to 53.9°C at Dému – show some variations for the three wells (see Appendix – Table A-1). These temperature differences are mainly due to pumping rates and cooling conditions within the hundreds of meters of casings: the wells are

not pumped continuously since the three of them are used for drinking water, with pumping fluctuations according to the hour of the day. The water temperatures of each well are consistent with the mean annual air temperature (MeteoFrance, 2020) and the mean geothermal gradient (ranging between 2.8 and 3.2°C/100 m) observed in the Aquitaine Basin (Housse and Maget, 1977).

Electric conductivity does not vary at Eauze whereas variations were observed of up to 10 % at Dému and 30 % at Nogaro 2 (See Appendix - Table A-1). These variations were due to changes in major element concentrations as described below.

Redox potential fluctuated greatly for the three wells, ranging from -300 to +150 mV. The values were mostly negative, which is coherent with reduced waters pumped at such depths. However, interpreting such fluctuations is tricky since this parameter is always difficult to measure, especially for moderately hot waters containing oxidized and reduced species (e.g. H₂S, sulfate, iron). The redox potential often corresponds to a global value impacted by different redox couples. We did not use this parameter for the interpretations.

Concerning pH, the values ranged between 7.1 and 7.8, which is coherent with waters whose pH is mainly controlled by calco-carbonic equilibrium with a relative low CO₂ partial pressure, around 10⁻² atm (André et al., 2005).

3.3.2 - Major elements

The major element concentrations are given in Appendix – Table A-2. Waters show a calcium bicarbonate facies, like most of the waters from the Eocene aquifer (André et al., 2005). They have a total dissolved solid concentration lower than 1000 mg.l⁻¹. Waters are generally at the equilibrium with quartz, calcite and a CO₂ partial pressure close to 10⁻² atm, whereas they are undersaturated with respect to sulfate minerals (see calculated saturation indices in Appendix – Table A-7). Since most of them have constant concentrations with time, Figure 6 focuses on sulfate alone. It shows how sulfate concentrations changed between January 2016 and March 2018 for the three investigated wells. For Nogaro 2, in addition to the nine sampling campaigns done by BRGM, the monthly sulfate concentration measurements organized by the Regional Health Agency are plotted on Figure 6. For Nogaro 2, three major sulfate concentration peaks were measured from January to April 2016 (with a maximum value of 34 mg.l⁻¹), from March to June 2017 (maximum value of 43 mg.l⁻¹) and from March to May 2018 (maximum value of 72 mg.l⁻¹). Sulfate concentrations for these peaks were 1.5 to 5 times the “reference sulfate concentration”, estimated at 10 mg.l⁻¹. Two very high sulfate concentrations were recorded during this period, the first one 3 January 2017 (with a value close to 300 mg SO₄.l⁻¹) and the

other 5 September 2017 (with a value of about $82 \text{ mg SO}_4\text{.l}^{-1}$). For the first one, the major increase of sulfate concentration is consistent with the water conductivity measurements, reflecting a substantial increase of water mineralization. However, as the major elements concentrations are not available for this sampling, it is impossible to confirm whether the value is consistent with the groundwater sulfate concentrations (analytical error, filtration issues in the samples, sulfate enriched particles, etc.) and with which other elements the sulfate would be correlated. It is also important to underline the brevity of this event since the sulfate concentration is about 18 mg.l^{-1} one week later.

The increased sulfate concentration in the Nogaro 2 waters seems to be correlated with increased calcium and magnesium concentrations (See Appendix - Table A-2). The $(\text{Ca}+\text{Mg})/\text{Cl}$ ratio clearly shows brief enrichment in Ca and Mg and the $(\text{Ca}+\text{Mg})/\text{SO}_4$ ratio decreases drastically, underlying the enrichment in both calcium and sulfate in the solution. The Dému well waters are the most stable, with no observable seasonal variations in major anion and cation concentrations. Each calculated ratio confirms that these concentrations remain stable over time (Figure 7). At the Eauze well, some variations in sulfate concentrations were observed in October 2016 and December 2017. The increase was a much smaller proportion than at Nogaro 2. The enrichment factor was about 1.2 compared to a base concentration of about $19\text{--}20 \text{ mg.l}^{-1}$. The ratio $(\text{Ca}+\text{Mg})/\text{SO}_4$ was weakly impacted by this increased sulfate concentration even though a slight decrease was observed in October 2016 and December 2017.

The piezometric level of the Nogaro 2 well (Figure 6) was anticorrelated with the increases in sulfate concentrations in late winter-early spring 2016, 2017 and 2018: during these periods, the sulfate concentration was at its highest when the piezometric level was at its lowest. These observations hold if we exclude the two inexplicable peaks measured in January 2017 and September 2017. For these two peaks, there are no anomalies on the piezometric data that could explain these change in concentrations.

While at Nogaro 2, peaks in sulfate concentration were anti-correlated with the piezometric level, for Eauze there is an almost positive correlation between sulfate concentration and piezometric variations. Much more regular sampling over time (as in Nogaro 2) would help to better define peak periods and thus better understand the links between these two variables (phase shift, lag or simultaneous signal relative to the other).

Figure 6 – Comparison between data measured during the project for wells: Nogaro 2 (left column), Eauze (centre column) and Dému (right column). First line of graphics: total sulfate concentration (dashed line) and static water level (black open symbols); second line of graphics: $\delta^{13}\text{C}$ content (dashed line) and static water level (black symbols); third line of graphics: $\delta^{34}\text{S}$ in sulfates (dashed line) and total sulfate concentration (full line); fourth line of graphics: $\delta^{11}\text{B}$ (dashed line) and total dissolved boron concentration (full black line).

3.3.3 - Trace elements

Variations in Al, B, Ba, F, Fe, Li, Mn, Sr and Br concentrations were measured in the waters of the three wells for all sampling campaigns (See Appendix - Table A-3).

Except the decrease in the strontium content in December 2017, for which we see no particular explanation (Perhaps a sampling or analytical problem?), strontium and barium concentrations were stable over time for the three wells (See Appendix - Table A-3). This stability in strontium is confirmed by isotope ^{87}Sr (see paragraph below). Barium concentration was low in waters

and does not seem to be controlled by barite (see Appendix – Table A-7). The saturation indices of barite calculated with PHREEQC code (Parkhurst and Appelo, 2013) were close to -0.8/-0.9 for Demu waters and to -0.20/-0.30 for Eauze waters. For Nogaro, because of the variation in sulfate concentration, the saturation indices range between 0.02 during sulfate peaks (like in March 2018) to -0.6 during low sulfate concentration phases (like in June 2017). Li concentrations do not vary with time, with values ranging between 8 and 15 $\mu\text{g.l}^{-1}$. Al concentrations do not exceed 2 $\mu\text{g.l}^{-1}$ and it seems controlled in some cases by equilibrium with montmorillonite or illite (See Appendix - Tables A-3 and A-7).

Finally, boron is the trace element that undergoes variations for all three wells. For the Nogaro 2 well, this variation is cyclical, with higher concentrations in April 2016, March 2017 and March 2018 (Figure 6). The concentration reaches as much as 80 $\mu\text{g.l}^{-1}$ and falls as low as close to 20 $\mu\text{g.l}^{-1}$. The variation of the concentration profile over time is similar to that of sulfate, with an increase in April 2016, March 2017 and March 2018, and somewhat of a plateau for the other measurements (around 20 $\mu\text{g.l}^{-1}$). For Eauze and Demu, fluctuations were observed but in a less extent than in Nogaro since the concentration varies generally between 42 and 48 $\mu\text{g.l}^{-1}$. There is no evidence of a cyclic evolution of the boron concentration. One value measured in January 2017 for both wells seems to fall outside of the range. It is lower by about 20 % than all the other measurements at these two wells but there is no coincidence with abnormal values of boron isotopes (See paragraph on ^{11}B isotope and Figure 6).

3.3.4 - Isotopes

Like for major and trace elements, we screened a wide range of stable and radioactive isotopes to identify potential tracers of the chemical mechanisms responsible for the observed variations. Isotopes of water (^2H and $\delta^{18}\text{O}$) and strontium isotopes (^{87}Sr) were measured during the first sampling campaigns (April 2016 to January 2017). They both had largely constant values and an absence of cyclic variations (see Appendix – Tables A-4 and A-5). Therefore, we focused more on interpreting the carbon, sulfate and boron isotopes. Isotopic compositions for stable isotopes of ^{34}S , ^{11}B and ^{13}C are reported in the usual δ -scale, with $\delta_{\text{sample}} (\text{‰}) = \{(R_{\text{sample}} / R_{\text{reference}}) - 1\} \times 1000$, where R is the ratio of the numbers of heavy and light isotopes.

- ^{13}C and ^{14}C in carbonates

Carbon-13 and carbon-14 in carbonates were analysed during the study of the three wells (See Appendix - Table A-6).

The waters of the Nogaro 2 show relatively stable carbon-14 activities, with small amplitude variations around 2 pMC (see also Andre et al., 2019). The apparent age of the waters is about -30,000 years BP according to Pearson and Hanshaw model (1970). $\delta^{13}\text{C}$ has cyclic variations, which seem to be correlated with the piezometric level over the period 2016-2018, with a high value for this parameter at the maximum piezometric level in September 2017 (Figure 6).

We also observed the same variations and correlation between $\delta^{13}\text{C}$ and the piezometric level for the two other wells. $\delta^{13}\text{C}$ variations reached about 3 ‰, as in Eauze, or only 1 ‰, as in Nogaro 2 and Dému. These variations could indicate that potential chemical reactions and/or water mixing processes could affect the $\delta^{13}\text{C}$ value in solution.

• ^{34}S in sulfates

For the Nogaro 2 well, variations in sulfur-34 follow the variations in total sulfur (Figure 6). Indeed, when the sulfate concentration is close to 10 mg.l^{-1} , the $\delta^{34}\text{S}$ value is close to +7.0 ‰. Though total sulfate concentration and sulfur-34 increase, variations in sulfur-34 are relatively small, as they do not exceed 2 ‰, even when the sulfate content is multiplied by 4, as in March 2018.

To sum up for the Nogaro 2 well, when the piezometric level is high, the sulfate content decreases and carbon-13 content tends to increase (anti-correlation between sulfate and carbon-13).

For the Eauze well, the relationship between the sulfate concentration and the $\delta^{34}\text{S}$ value is much less pronounced than at Nogaro 2. The small seasonal variations in sulfate concentrations (lower than 5 mg.l^{-1}) correlated with piezometric fluctuations and the small amplitudes recorded for $\delta^{34}\text{S}$ do not allow us to infer an accurate correlation.

For the Dému well, the variations in total sulfur concentration are very low (less than 1 mg.l^{-1}) and confirmed by the absence of $\delta^{34}\text{S}$ variations. We note that the $\delta^{34}\text{S}$ was quite high (about +17 ‰), much higher than at the Eauze well (about +9.5 ‰), which seems to indicate a different origin for sulfur in solution.

• ^{11}B

The waters of the Nogaro 2 well show a cyclic and seasonal variation for $\delta^{11}\text{B}$ value. When the boron content increased in water (April 2016, March 2017 and March 2018), the $\delta^{11}\text{B}$ significantly decreased to reach the order of +10 ‰ (Figure 6). When boron levels were lower (on the order of 25 mg.l^{-1}), $\delta^{11}\text{B}$ increased to +18 ‰. This finding seems to indicate a mix of the initial water with a ^{11}B -depleted solution. For Eauze well, the variations of $\delta^{11}\text{B}$ values are limited

to 2 ‰. The relation between $\delta^{11}\text{B}$ and the total boron concentration is less remarkable than for Nogaro 2 even if a cyclic variation can be supposed between mid 2016 and mid-2017. For Demu well, $\delta^{11}\text{B}$ values are stable with time.

3.4 – Geochemical modelling

Analyses of major elements, trace elements and isotopes in the Nogaro 2 well waters confirm cyclic variations in sulfate concentrations already observed in the chemical time series. These sulfate concentration increases are correlated with increases in boron levels and anti-correlated with static water level and ^{13}C content in the waters.

Based on sulfur-34 isotopic data, sulfate enrichment seems to come from a source with a $\delta^{34}\text{S}$ greater than the base value in the reservoir (estimated here at +6.5 ‰ CDT). For boron, the increase in its total concentration coincides with a decrease of $\delta^{11}\text{B}$. In this case, the source of boron responsible for the increase in concentration has a lower $\delta^{11}\text{B}$ value than the base value in the reservoir (estimated at +17.5 ‰).

To interpret these field results, simulations with the PHREEQC software (Parkhurst and Appelo, 2013) were carried out with the *iso.dat* database released with the code. Since boron-11 is not in the database, the new species was implemented: parameters used to calculate the specific volume of H_3BO_3 are from SUPCRT92 (Johnson et al., 1992), whereas the standard value (4.04362 ‰) for $^{11}\text{B}/^{10}\text{B}$ is from Catanzaro et al. (1970). This software is particularly used in geosciences because of its adaptability to geochemical problems of varying complexity. It offers, among other things, the possibility of making geochemical calculations on isotopes, taking into account the ratios of different species. The simulations consisted of reproducing variations in sulfate and boron levels in the Nogaro 2 waters since this well presents the most significant variations. The main objective was to be able to establish more precisely the values of $\delta^{34}\text{S}$ (in sulfate) and $\delta^{11}\text{B}$ (for boron) that participate in these cyclical concentration variations. All modelling was done in a batch system in which water mixtures (in the form of additions) were simulated (Figure 7).

As shown in Figure 8, the initial water used for the simulations corresponded to the Nogaro 2 water analysed in April 2016, the beginning of quarterly samples, with an initial sulfate concentration of around 24 mg.l^{-1} (and $\delta^{34}\text{S} = +8.0$ ‰) and a boron content of 44 $\mu\text{g.l}^{-1}$ (and $\delta^{11}\text{B} = +11.5$ ‰). The numerical simulations consisted of modelling two cycles, each comprising two stages, namely a dilution phase (by mixing with a sulfate- and boron-depleted fluid) and a concentration phase (by mixing with a fluid enriched with sulfate and boron). The durations of

the dilution and concentration phases are consistent between simulations and on-site measurements.

A first dilution phase lasting 180 days was then simulated. The dilutions were done by successive additions of a Fluid 1 whose sulfate and boron contents were respectively 10 mg.l⁻¹ and

24 µg.l⁻¹ (Figure 7). If Fluid 1 was assumed to be initially in equilibrium with calcite and to have a CO₂ partial pressure of 10⁻² atm, this equilibrium is not imposed during all the dilution phase (no constraints on the calco-carbonic equilibrium). Initial pH of water is close to 7.35 and Eh = 0 mV (consistent with measured values – see Appendix – Table A-1). The mixing consisted in adding Fluid 1 by 5 % in volume every day. The resulting sulfate and boron concentrations fell to 10 mg.l⁻¹ for sulfate (and δ³⁴S = +6.5 ‰) and 24 µg.l⁻¹ for boron (and δ¹¹B = 17.5 ‰). These modelled variations were consistent with the measured groundwater concentrations, simulating both the decrease and the plateau of sulfate concentration (Figure 8).

Figure 7 – Conceptual model of the geochemical modelling of the dilution and concentration stages impacting the Nogaro 2 waters.

Next, a first concentration phase was simulated over a period of 100 days (Figure 7). This phase was shorter since it only represented the increased sulfate concentration whereas the previous included both the decreasing sulfate concentration and the plateau. The added solution (called Fluid 2) had sulfate and boron concentrations of 1500 mg.l^{-1} and 1620 µg.l^{-1} , respectively. Fluid 2 was assumed to be in equilibrium with calcite to impose the carbon content corresponding to a CO_2 partial pressure around $10^{-2.1}$ atm. The isotope values were $+10.0 \text{ ‰}$ and -3.2 ‰ for sulfur of sulfate and boron, respectively. These concentration and isotope values were calibrated to best represent the measured data. Among all the potential solutions, we took chose to consider the mixing fluid in equilibrium with gypsum, at the temperature of the reservoir. Since this water-mineral equilibrium fixes the sulfate concentration, it was then possible to determine both the amount of the added fluid and the $\delta^{34}S$ value of sulfate to match the measured data. The boron concentration was characterized according to the amount of fluid added and finally, the boron isotope value was defined. Fluid 2 was added in 0.0075 ‰ volume increments every day, which is a far lower percentage than during dilution 1. That discrepancy

595 between mixing proportions is addressed in section 4. The consequence of these successive
 596 additions was an increase in sulfate and boron concentrations and $\delta^{34}\text{S}$ and a decrease in $\delta^{11}\text{B}$
 597 over time (Figure 8). The maximum peak values are respected.

Figure 8 – Variations with time of (a) total sulfate concentration and $\delta^{34}\text{S}$; (b) total boron concentration and $\delta^{11}\text{B}$ in the Nogaro 2 water. Dilutions and concentration correspond to the stages defined in Figure 7. Symbols are data measured during this study.

The second cycle began with a dilution phase, significantly longer than the dilution during phase 1 (300 days), so we reached the base values in the aquifer. Then a second concentration phase over 100 days followed. During this phase, the maximum boron and sulfate concentrations were much higher than in the first concentration phase, whereas the final values of the isotopic

values were similar to the initial values. This suggests that the mix is more complex than in the first case and another source of sulfate and boron is required. A new “recharge” solution (called Fluid 3) had to be defined, with sulfate concentrations on the order of 1000 mg.l^{-1} and a boron concentration of about $1620 \text{ }\mu\text{g.l}^{-1}$. Fluid 3 was assumed to be in equilibrium with calcite to impose the carbon content corresponding to a CO_2 partial pressure around $10^{-2.1} \text{ atm}$. Isotope values for boron and sulfur of sulfate were $+12 \text{ ‰}$ and $+6.5 \text{ ‰ CDT}$, respectively. As previously, these values were numerically determined to best represent the measured data. Each Fluid, 2 and 3, was added in increments of 0.015 ‰ by volume every day. These additions increased the sulfate and boron contents in the mixture (to reach respectively 46 mg.l^{-1} and $80 \text{ }\mu\text{g.l}^{-1}$), while respecting the values of measured isotopic ratios.

These geochemical simulations allowed us to estimate the concentrations and isotope compositions of waters that can mix with the Eocene aquifer groundwater to explain the cyclical variations, especially in sulfate and boron. These simulations relate to “classic” variations in sulfate concentration, but they do not represent exceptional variations, such as the sulfate concentration of 300 mg.l^{-1} recorded in January 2017 as part of the Health monitoring of Nogaro 2 waters. According to these simulations, the “recharge” water used during the concentration phases had significantly higher sulfate and boron content than the waters of the aquifer.

4 – Discussion

This study confirms the cyclic variations of sulfur concentration observed during recent years in the waters of Nogaro 2 well and also, but with a lower amplitude, at Eauze. There are no variations in Dému, which is farther from Nogaro 2. As discussed in André et al. (2002), sulfur in solution can have different origins. Some can come from precipitation (since water has meteoric origins), from gypsum dissolution, or from the oxidation of reduced sulfur minerals like pyrite. Our study does not describe explicitly the potential chemical mechanism responsible for these variations. But the numerical approach used in this study allows us to introduce some hypotheses to explain these cyclic variations.

First, it seems that the Eocene groundwater (named Fluid 1) close to the Nogaro 2 well has a “base” sulfate concentration value of 10 mg.l^{-1} and a $\delta^{34}\text{S}$ of $+6.5 \text{ ‰ CDT}$, which could reflect various origins for sulfur (mainly gypsum dissolution and a small part from pyrite oxidation according to André et al., 2002). The cyclic concentration increases could be due to mixing with a sulfate-enriched fluid. Many numerical solutions are possible since the chemical compositions of the added fluids and the proportion of the mix are unknown. We selected the option to consider the mixing fluid in equilibrium with gypsum at the temperature of the reservoir for two

reasons: first gypsum has been identified in the molasse horizon, at the top of the aquifer. Molasse sediments are known to be a low-permeable media, fully saturated with water. Because of the long contact time between water and the rock formations, we can assume that the solution and minerals are in equilibrium. Secondly, since $\delta^{34}\text{S}$ varies by relatively low amounts, the added fluid must have a $\delta^{34}\text{S}$ value close to the one in the reservoir. The geochemical modelling shows that $\delta^{34}\text{S}$ of the mixed fluid is close to +10 ‰ CDT, which seems to be in coherence with an evaporitic origin of sulfate, like gypsum dissolution (André et al., 2002; Négrel et al., 2009 ; Malcuit, 2012). That is also consistent with values given in the literature (Claypool et al., 1980) and it is not very far from the +12.4 ‰ CDT measured on gypsum crystals sampled in the molasse unit of the Aquitaine Basin (André et al., 2002). For Fluid 3, the ^{34}S selected had the same value than in the reservoir, which could indicate a higher proportion of sulfur coming from sulfide minerals. $\delta^{18}\text{O}_{\text{SO}_4}$ for Nogaro 2 are close to + 10.5 to 14.6 ‰ with low variations with time, which is consistent with the mixing of sulfate having $\delta^{18}\text{O}_{\text{SO}_4}$ value close to 14.9 ‰, like the one measured on the gypsum from molasse by André et al. (2002).

For boron, the proposed $\delta^{11}\text{B}$ values were on the order of -3.2 ‰ for Fluid 2 and +12 ‰ for Fluid 3. During this study, a new lab experiment was conducted to identify the potential origin of this boron. We did not identify any boron-bearing mineral formally among the primary minerals in the aquifer or in the molasse aquitard. Therefore, rock samples from cuttings and cores sampled at the top of the aquifer (sampled in a neighbouring well drilled recently) were analysed (XRD analysis, micro-probe and more) and laboratory tests were carried out to dissolve these samples. The experiments consisted of dissolving 10 g of crushed material (cuttings) in 100 mL of water and measuring, after one month, the concentrations of major and trace elements in the aqueous solution. The experiments were carried out at a temperature of 38°C and atmospheric pressure. We studied twelve samples from different geological horizons (molasse aquitard and Lussagnet sands aquifer). After one month, analysis of the water phase showed quite large disparities between the samples. Essentially, at a depth of 610 m, i.e. in the heart of the Lussagnet Sands aquifer, for which the depth ranges between 550 m and 650 m, the $\delta^{11}\text{B}$ value is similar to that used for the geochemical simulations (Figure 9).

Figure 9 – Boron concentration and $\delta^{11}\text{B}$ values in the aqueous phase resulting from the lab experiments on the dissolution of cuttings sampled at different depths from a well located near the Nogaro 2 well. The top of the Lussagnet sands aquifer starts at a depth of 550 m. The grey-shaded zone indicates the molasse unit, which overlies the Lussagnet sand aquifer.

Because of the low $\delta^{11}\text{B}$ values identified in the Lussagnet sand aquifer (between -3.15 and +2.00 ‰), it could be inferred that boron may come from the alteration of silicates. Indeed, the alteration of silicate minerals by waters gives low $\delta^{11}\text{B}$ values, ranging between -10 and +10 ‰ (Barth, 2000; Pennisi et al., 2000; Casanova et al., 2002; Négrel et al., 2002 ; Lemarchand and Gaillardet, 2006; Millot et al., 2007; Clauer et al., 2018). As shown in Figure 10, this value is also consistent with the end-member of silicates identified by Négrel et al. (2009). The boron content of Fluid 2 could be explained by the alteration of the silicate sands of Lussagnet and this fluid could mix with the water initially present in the Baliros sands Formation exploited at Nogaro 2. During the second modelled cycle, the added Fluid 3 had a slightly higher $\delta^{11}\text{B}$ value at +12 ‰. Literature data show that the $\delta^{11}\text{B}$ value of evaporites ranges between +15 up to +35 ‰ according to the evaporation level of the brines (Swihart et al., 1986; Vengosh et al., 1992; Liu et al., 2000). Figure 10 confirms the potential evaporitic origin of this boron. This value is also close to the mixing line between clays and evaporites calculated by Négrel et al. (2009). Therefore we assumed that during this second cycle the boron dissolved in Fluid 3 comes from the molassic aquitard, where evaporitic sediments have been identified. Our study seems to confirm that the top of the productive part of the aquifer and potentially the molasse could be two sources of elements that could feed the exploited part of the aquifer (Baliros sands Formation) according to pressure conditions in the reservoir.

Figure 10 – Variations of $\delta^{11}\text{B}$ versus the boron concentrations in the waters from Nogaro 2, Eauze and Demu. Fluids 1, 2 and 3 correspond to the fluids used for the geochemical calculations. End-members (carbonate, evaporite, clays and silicates) and mixing lines are issued from Négrel et al. (2009).

These potential contributions from the overlying molassic horizons had already been assumed by André (2002). Indeed, the upper formation contains gypsum, identified in cuttings. The upper aquitard would therefore be a sulfate reserve and sulphur could be transferred into the reservoir. We did not investigate the exact transfer mechanisms responsible for the variations of certain elements without impacting the concentration of the other elements in detail. Convection and/or diffusion are potential processes to explain the seasonal fluctuations (Atteia et al., 2005). A sensitive hydrogeologic analysis coupled with the chemical hypothesis defined in this study could help to better characterize them. We can however observe from these chemical simulations that the mixing proportions differ between the dilution and the concentration phases. During the dilution phase, the amounts added every day are relatively high, whereas during the concentration phase the daily amounts are very low. Interpreting this difference is tricky since only the mixing processes are used in this study without any hydraulic constraints. However, one interpretation could be that two processes are in competition. On one side, we can suppose that the dilution phase could be assimilated to a hydraulic process with a high renewal rate of the aquifer water (or a low resident time). On the other side, the concentration phase could correspond to a process with a longer resident time favouring the mixing of the aquifer water with a concentrated solution. Though this study helps us to understanding of the chemical

variations, it also underlines that the aquifer should not be considered as a single horizon and that local structural and hydrogeological heterogeneities may explain vertical differentiation in the aquifer.

This does raise a question regarding these geochemical simulations. Since it is assumed that the mixed concentrated fluids are enriched in sulfate and in equilibrium with gypsum, the calcium concentration in solution should increase (Figure 11). But the sulfate increase does not follow the same trend as that of calcium (and/or magnesium). Figure 16 clearly shows that the measured calcium concentrations are lower than the calculated values. This could mean that calcium (and/or magnesium) is controlled by another mechanism.

Figure 11 – Variation in calcium concentration. Curves are results of two simulations (with and without maintaining calcite equilibrium) whereas symbols are calcium concentration measured during quarterly samplings

During the addition of concentrated fluids (Fluids 2 and 3), calcium is added in a solution (Fluid 1) initially in equilibrium with calcite. Then the saturation indices of calcite increases and the carbonate mineral becomes over-saturated, whereas the gypsum remains under-saturated (Figure 12).

Figure 12 – Variation of the simulated saturation indices of calcite (left) and gypsum (right) during the mixing of Fluid 1 with Fluids 2 and 3. Lines are results of simulations whereas symbols are saturation indices calculated from chemical compositions of waters measured during quarterly samplings

Because of calcium's high kinetic reactivity, this mineral is often assumed to be in equilibrium with the solution. A new geochemical run was then performed assuming calcite equilibrium and imposing a CO_2 partial pressure of 10^{-2} atm during the mixing. Under this condition, the calcium concentration in solution is controlled by calco-carbonic equilibrium and secondary calcite is precipitating in the reservoir. By imposing this constraint to the chemical model, Figure 11 shows that the calcite precipitation controls and decreases the calcium content in the solution and numerical results are more coherent with measured data. Moreover, it is admitted that light isotopes concentrate in the least dense phase, i.e. in the liquid phase instead of the solid phase (Allègre and Michard, 1973 ; Hoefs, 1997; Couchoud, 2008). Consequently, ^{13}C precipitates in priority under calcite, the solution is enriched with the light isotope and the $\delta^{13}\text{C}$ value decreases. It corresponds to the observations in Nogaro 2 since sulfate concentration and $\delta^{13}\text{C}$ are anticorrelated : during the concentration phase, sulfate increases, calcite precipitates and $\delta^{13}\text{C}$ decreases (Figure 6).

6 - Conclusions

The 2-year monitoring of concentrations of major and trace elements and stable and radioactive isotopes confirms the cyclical and seasonal variations in sulfate (between 10 and 40 mg.l^{-1}) as well as variations in sulfur-34 of sulfate (between +6.5 and +8.5 ‰) at the Nogaro 2 well. Though sulfate variations were known, additional variations in boron (total concentration and isotope) and carbon (^{13}C and ^{14}C) contents have been also highlighted by these new campaigns. However, given the very small activity in ^{14}C , it has been difficult to conclude on the origin of the measured variations (geochemical processes, water mixtures, uncertainties related

to samples and analyses, for instance). Further study of this parameter could be considered, in particular to define the uncertainty associated with abstractions and analyses.

In terms of interpretations, geochemical modelling based on water mixtures has revealed concentration processes in relation with mixtures of sulfate and boron-enriched waters and dilution stages, corresponding to mixes between water from the reservoir and water coming from other origins. While the geochemical models make it possible to properly restore both sulfate and boron concentrations in the water (as well as the $\delta^{34}\text{S}$ and $\delta^{11}\text{B}$ values), this approach must be considered exploratory in order to propose mixing end-members. Ideally, this study should be supplemented by a more accurate analysis of the minerals present at the site (in the reservoir and in the layers both beneath and overlying). This mineral study would allow us to define more precisely the phases bearing sulfur and boron and to characterize them (especially in terms of isotopic content). The geological and mineralogy study we carried out confirms that in the Nogaro 2 zone sulfur may have two origins: gypsum, which is generally present in the molassic aquitard, and pyrite, present in the reservoir. However, it has not been possible to characterize the sulfur of these pyrites precisely, especially in terms of isotope. The small amount of pyrite and the coating of the grains by quartz deposits prevented any purification and isotopic analysis. However, these characterizations would be necessary to determine the origin of sulfate in the waters, especially during peak phases.

This study therefore provides important elements for understanding the processes responsible for variations in sulfate concentration in waters from Nogaro 2. It appears that these variations are probably related to coupled hydrodynamic and geochemical processes and that only a finer characterization of the end-members (origin of sulfur and boron, phases bearing these elements, stock/reserve of these elements, etc.) and exchanges between the different parts of the reservoir (and the aquitards) will help to understand the mechanisms occurring at this borehole. A coupled hydrogeological and chemical approach will also be needed to better understand these processes.

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Appendix

Table A-1: Physico-chemical parameters of the waters measured on site

Well	Date	Conductivity ($\mu\text{S/cm}$)	pH	T°C	O ₂ (%)	O ₂ (mg/l)	Eh _{measured} (mV)	Eh _{corrected} / ENH (mV)
NOGARO 2	19/04/2016	325	7.2	51	42	2.35	-209	-19
	18/07/2016	296	7.13	50.9	3.3	0.19	-284	-94
	04/10/2016				Not measured			
	10/01/2017	347	7.17	48.7		0.65	-199	-7
	21/03/2017	384	7.75	42.0	72.8	4.84	-129	68
	27/06/2017	328	7.55	49.1	20.2	1.18	-93	98
	03/10/2017	296	7.48	47.2	39.7	2.29	-112	81
	14/12/2017	338	7.71	46	67.8	4.04	-122	72
	01/03/2018	365	7.85	44.5	82.5	4.77	94	289
EAUZE	19/04/2016	308	7.32	36.4	3.2	0.21	-169	30
	18/07/2016	307	7.14	36.4	5.9	0.4	-215	-14
	04/10/2016				Not measured			
	10/01/2017	333	7.4	31.5		0.53	-165	39
	21/03/2017	325	7.73	31.8	48.2	3.76	-80	124
	27/06/2017	330	7.75	35.5	58.8	4.01	53	254
	03/10/2017	319	7.23	27.8	53.9	4.2	135	342
	14/12/2017	326	7.65	32.0	38.5	2.78	18.4	222
	01/03/2018	316	7.5	34.9	40.5	2.71	95	297
DEMU	19/04/2016	298	7.38	48.1	29.1	1.83	-187	5
	18/07/2016	293	7.14	53.9	56.8	3.10	-219	-31
	04/10/2016				Not measured			
	10/01/2017	326	7.25	53.2		0.10	-198	-10
	21/03/2017	323	7.44	46.6	14.3	0.91	-145	48
	27/06/2017	325	7.46	48.0	15.8	0.88	-125	67
	03/10/2017	296	7.21	53.0	3.0	0.16	-211	-22
	14/12/2017	326	7.45	49.0	15.0	0.92	-173	18
	01/03/2018	462	7.25	48.0	12.8	0.73	52	244

Table A-2: Chemical composition of major elements (Concentrations in mg.l⁻¹) and calculated molar ratio

Date	Well	Ca	Mg	K	Na	SO ₄	HCO ₃	Cl	SiO ₂	Ca + Mg	Ca + Mg	Na + K
										SO ₄	Cl	Cl
Apr.-16	NOGARO 2	39.2	5.2	5.2	14.5	24.2	155	7.5	23.1	4.74	11.30	3.61
	EAUZE	35.4	5.1	5.7	15.7	19.5	153	6.5	17.4	5.39	11.96	4.52
	DEMU	32.2	4.5	5.8	17.5	6.8	160	7.9	24.5	13.98	8.90	4.09
Jul.-16	NOGARO 2	37.4	4.5	5.3	14.1	10.1	153	7.6	24.3	10.65	10.46	3.50
	EAUZE	36.0	5.3	5.8	16.4	18.8	150	6.5	18.9	5.71	12.21	4.70
	DEMU	33.7	4.6	5.9	18.2	7.0	157	7.9	26.3	14.15	9.27	4.23
Oct.-16	NOGARO 2	37	4.6	5.7	14.6	10.5	153	7.7	24.3	10.19	10.27	3.60
	EAUZE	36.8	5.6	6.2	17.0	23.9	149	6.6	18.9	4.62	12.38	4.83
	DEMU	32.7	4.6	6.3	18.9	6.3	158	8.0	25.7	15.34	8.94	4.36
Jan.-17	NOGARO 2	39.1	4.9	5.5	15.2	17.9	152	7.8	24.1	6.32	10.73	3.65
	EAUZE	35.8	5.4	6.0	16.6	19.8	148	6.7	18.4	5.42	11.84	4.64
	DEMU	33.2	4.6	6.0	19.0	6.8	155	8.1	25.8	14.39	8.93	4.29
Mar-17	NOGARO 2	37.2	4.7	5.6	14.9	14.9	154	7.5	24.6	7.24	10.64	3.74
	EAUZE	35.3	5.3	5.9	16.6	19.8	151	6.3	18.8	5.34	12.40	4.92
	DEMU	33.1	4.7	6.0	18.6	6.4	159	7.8	26.5	15.31	9.29	4.38
Jun-17	NOGARO 2	38.8	4.4	5.5	14.5	10.7	149	7.6	25.4	10.33	10.75	3.60
	EAUZE	37.3	5.1	6.0	16.7	19.7	145	6.5	19.5	5.57	12.48	4.80
	DEMU	34.5	4.4	6.2	18.7	6.9	152	7.9	26.8	14.52	9.38	4.37
Oct.-17	NOGARO 2	37.3	4.5	5.2	13.5	10.6	154	7.7	23.1	10.12	10.31	3.32
	EAUZE	36.6	5.3	5.8	15.6	20.2	151	6.7	17.3	5.39	12.01	4.38
	DEMU	33.4	4.6	5.9	17.5	6.6	158	8.1	24.5	14.90	8.98	4.00
Dec.-17	NOGARO 2	37.6	4.5	5.6	14.1	11.2	153	7.7	24.3	9.64	10.38	3.49
	EAUZE	37.1	5.3	6.0	16.4	23.2	149	6.5	19.1	4.74	12.51	4.73
	DEMU	33.2	4.6	6.1	18.0	6.4	157	8.0	26.3	15.29	9.05	4.17
Mar-18	NOGARO 2	44.8	6.3	5.7	16.3	46.7	154	7.6	24.9	2.84	12.89	3.99
	EAUZE	35.4	5.3	5.9	15.9	20.4	151	6.7	19.0	5.19	11.69	4.46
	DEMU	32.8	4.5	6.0	17.8	7.1	158	8.1	26.5	13.59	8.81	4.06

Table A-3: Concentrations of trace elements

Date	Well	Al µg/l	B µg/l	Ba µg/l	F mg/l	Fe mg/l	Li µg/l	Mn µg/l	Sr µg/l	Br µg/l
Apr.-16	NOGARO 2	1.13	43.9	120	0.2	0.208	11.20	7.8	496	< LQ
	EAUZE	0.66	42.2	95	0.2	0.108	10.20	12.3	466	81
	DEMU	1.61	42.3	113	0.2	0.232	13.80	13.4	457	54
Jul.-16	NOGARO 2	1.02	24.3	111	0.2	0.225	9.38	6.4	438	<50
	EAUZE	0.58	42.5	97	0.2	0.104	10.20	11.9	475	<50
	DEMU	1.40	43.7	112	0.2	0.210	13.90	12.1	468	<50
Oct.-16	NOGARO 2	1.92	24.7	104	0.2	0.184	9.88	5.2	455	145
	EAUZE	1.16	47.6	90	0.2	0.109	11.60	12.6	480	208
	DEMU	1.05	47.2	105	0.2	0.265	15.20	19.4	457	99
Jan.-17	NOGARO 2	0.6	29.3	91.4	0.2	0.309	8.27	10.9	436	< LQ
	EAUZE	0.56	35.4	80.5	0.2	0.349	8.97	13.9	445	< LQ
	DEMU	0.73	33.3	93.1	0.2	0.343	10.9	12.7	434	< LQ
Mar.-17	NOGARO 2	NM	33.7	109	0.2	0.192	11.2	6.63	442	157
	EAUZE	NM	45.1	91.7	0.2	0.191	11.2	17.1	458	147
	DEMU	NM	46.1	109	0.2	0.122	15.3	18.6	440	129
Jun.-17	NOGARO 2	1.52	25.4	110.00	0.1	< LQ	9.86	6.4	433	29.0
	EAUZE	1.90	46.6	111.00	0.2	0.173	14.5	12.6	473	34.6
	DEMU	0.61	46.0	95.60	0.2	< LQ	10.5	12.4	470	27.1
Oct.-17	NOGARO 2	1.52	22.6	112.00	0.1	0.2	10.1	5.52	414	67.1
	EAUZE	1.47	41.5	96.20	0.2	< 0.02	10.7	12.6	462	44.7
	DEMU	1.46	42.3	113.00	0.2	0.376	14.4	17.2	450	49.0
Dec.-17	NOGARO 2	1.90	25.7	105.00	0.1	0.05	10.4	5.51	309	28.6
	EAUZE	0.90	43.4	86.60	0.2	< 0.02	10.5	11.5	330	24.8
	DEMU	1.43	42.5	106.00	0.2	0.352	14.0	15.0	319	34.0
Mar.-18	NOGARO 2	1.95	79.5	101.87	0.2	0.027	14.71	8.83	570	30.9
	EAUZE	0.69	44.3	93.93	0.2	0.036	9.51	12.13	455	28.4
	DEMU	2.00	45.1	109.36	0.2	0.214	14.68	11.84	449	36.4

< LQ : lower to quantification limits ; NM : not measured

Table A-4: ^2H and $\delta^{18}\text{O}$ of water molecule

	δD (‰ vs SMOW) (+/- 0.8 ‰)				$\delta^{18}\text{O}$ (‰ vs SMOW) (+/- 0.1 ‰)			
	Apr-16	Jul-16	Oct-16	Jan-17	Apr-16	Jul-16	Oct-16	Jan-17
NOGARO 2	-54.3	-54.7	-54.3	-54.5	-8.5	-8.5	-8.5	-8.5
EAUZE	-52.6	-52.5	-52.7	-53.0	-8.3	-8.3	-8.2	-8.4
DEMU	-52.7	-52.7	-52.6	-53.0	-8.3	-8.3	-8.4	-8.3

Table A-5: $^{87}\text{Sr}/^{86}\text{Sr}$ ratio

	Apr-16		Jul-16		Oct-16		Jan-17	
	$^{87}\text{Sr}/^{86}\text{Sr}$	$2\sigma(\text{m})$	$^{87}\text{Sr}/^{86}\text{Sr}$	$2\sigma(\text{m})$	$^{87}\text{Sr}/^{86}\text{Sr}$	$2\sigma(\text{m})$	$^{87}\text{Sr}/^{86}\text{Sr}$	$2\sigma(\text{m})$
NOGARO 2	0.708704	0.000008	0.708712	0.000007	0.708717	0.000007	0.708710	0.000006
EAUZE	0.708698	0.000005	0.708700	0.000006	0.708701	0.000008	0.708700	0.000008
DEMU	0.708781	0.000006	0.708778	0.000009	0.708778	0.000006	0.708774	0.000008

Table A-6: Isotope concentrations

Date	Well	$\delta^{13}\text{C}$ ‰ vs PDB (± 0.1)	Act. ^{14}C (pMC)	$\delta^{11}\text{B}$ ‰ (Try 1)	$\pm 2\sigma(\text{m})$ ‰	$\delta^{11}\text{B}$ ‰ (Try 2)	$\pm 2\sigma(\text{m})$ ‰	$\delta^{34}\text{S}$ (SO_4) ‰ vs CDT (± 0.3 ‰)	$\delta^{18}\text{O}$ (SO_4) ‰ vs SMOW (± 0.5 ‰)
Apr.-16	NOGARO 2	-14.60	1.8 ± 0.1	11.67	0.06	11.49	0.11	8.00	13.5
	EAUZE	-14.60	1.5 ± 0.1	13.97	0.06	13.58	0.08	9.30	12.8
	DEMU	-15.30	3.2 ± 0.1	17.94	0.04	17.70	0.09	15.90	10.2
Jul.-16	NOGARO 2	-14.80	1.5 ± 0.1	17.57	0.07	17.63	0.09	6.80	$10.6 (\pm 0.6 \text{ ‰})$
	EAUZE	-14.30	0.5 ± 0.1	15.23	0.09	15.32	0.06	9.50	$14.1 (\pm 0.6 \text{ ‰})$
	DEMU	-15.50	1.1 ± 0.1	19.47	0.06	19.44	0.11	15.80	10.7
Oct.-16	NOGARO 2	-14.30	2.0 ± 0.1	15.45	0.05	15.83	0.08	$6.80 (\pm 0.4 \text{ ‰})$	10.4
	EAUZE	-14.00	0.1 ± 0.1	13.68	0.06	13.31	0.08	9.50	12.3
	DEMU	-15.30	1.5 ± 0.1	19.03	0.05	18.81	0.08	18.40	11.5
Jan.-17	NOGARO 2	-15.00	1.3 ± 0.0	10.95	0.09	11.23	0.05	8.10	13.0
	EAUZE	-14.80	< 0.44	14.65	0.08	14.28	0.06	10.00	13.0
	DEMU	-15.70	0.9 ± 0.0	18.67	0.05	18.61	0.07	17.90	11.0
Mar.-17	NOGARO 2	-15.20	1.7 ± 0.0	9.64	0.06	9.89	0.04	7.00	11.00
	EAUZE	-15.30	0.8 ± 0.0	12.80	0.05	12.90	0.04	10.00	12.60
	DEMU	-15.50	0.8 ± 0.0	19.02	0.17	18.84	0.10	17.80	11.40
Jun.-17	NOGARO 2	-14.60	1.9 ± 0.0	16.90	0.18	17.09	0.14	6.50	10.50
	EAUZE	-13.70	1.4 ± 0.0	14.86	0.27	14.66	0.09	9.60	12.60
	DEMU	-15.30	1.1 ± 0.0	18.80	0.07	18.94	0.06	17.10	11.20
Oct-17	NOGARO 2	-13.00	1.5 ± 0.0	17.16	0.03	17.17	0.05	6.60	11.10
	EAUZE	-13.80	0.9 ± 0.0	14.63	0.03	14.25	0.05	9.80	13.30
	DEMU	-14.90	0.9 ± 0.0	18.65	0.04	18.82	0.04	20.00	11.80
Dec-17	NOGARO 2	-13.80	2.2 ± 0.0	15.94	0.04	15.97	0.05	6.70	11.20
	EAUZE	-11.30	1.7 ± 0.0	14.18	0.05	13.98	0.06	9.20	13.40
	DEMU	-14.50	1.1 ± 0.0	18.91	0.10	18.77	0.18	17.80	11.40
Mar-18	NOGARO 2	-13.80	1.5 ± 0.0	7.98	0.05	7.57	0.18	8.20	14.60
	EAUZE	-14.00	0.6 ± 0.0	13.89	0.06	13.76	0.23	9.20	13.20
	DEMU	-15.10	1.0 ± 0.0	18.62	0.05	18.17	0.04	17.90	11.50

< LQ : lower to quantification limits ; NM : not measured

Table A-7: Saturation indices of minerals calculated with PHREEQC program and THERMODDEM database (Blanc et al., 2012)

Well	Date	Calcite	Dolomite	Gypsum	Anhydrite	Barite	Strontianite	Quartz	Chalcedony	Montmorillonite (MgCa)	Illite (Al)	Log pCO _{2(g)}
NOGARO 2	19/04/2016	-0.05	-0.40	-2.27	-2.19	-0.24	-1.66	0.01	-0.25	-0.72	-0.99	-1.84
	18/07/2016	-0.14	-0.61	-2.65	-2.57	-0.64	-1.78	0.04	-0.22	-0.68	-0.89	-1.77
	04/10/2016	-0.28	-0.90	-2.64	-2.57	-0.64	-1.90	0.05	-0.21	-0.09	0.26	-1.65
	10/01/2017	-0.12	-0.58	-2.40	-2.34	-0.46	-1.77	0.06	-0.21	-0.93	-1.36	-1.83
	21/03/2017	0.34	0.30	-2.52	-2.53	-0.39	-1.24	0.14	-0.13	NC	NC	-2.46
	27/06/2017	0.25	0.13	-2.62	-2.56	-0.61	-1.40	0.07	-0.19	-0.44	-0.90	-2.22
	03/10/2017	0.16	-0.05	-2.64	-2.60	-0.58	-1.49	0.05	-0.21	-0.42	-0.68	-2.15
	14/12/2017	0.36	0.35	-2.62	-2.59	-0.57	-1.40	0.09	-0.18	-0.23	-0.61	-2.39
	01/03/2018	0.52	0.73	-1.98	-1.96	0.02	-1.04	0.11	-0.15	-0.11	-0.65	-2.55
EAUZE	19/04/2016	-0.19	-0.74	-2.44	-2.50	-0.27	-1.69	0.06	-0.22	-0.60	-0.39	-2.06
	18/07/2016	-0.37	-1.09	-2.44	-2.50	-0.27	-1.87	0.10	-0.18	-0.45	-0.12	-1.89
	04/10/2016	-0.03	-0.41	-2.34	-2.42	-0.19	-1.53	0.11	-0.17	-0.03	0.30	-2.27
	10/01/2017	-0.19	-0.77	-2.43	-2.54	-0.27	-1.69	0.14	-0.14	-0.27	0.01	-2.19
	21/03/2017	0.14	-0.10	-2.44	-2.54	-0.21	-1.34	0.14	-0.13	NC	NC	-2.51
	27/06/2017	0.22	0.05	-2.42	-2.48	-0.19	-1.29	0.12	-0.16	0.18	0.40	-2.53
	03/10/2017	-0.40	-1.23	-2.42	-2.56	-0.13	-1.87	0.16	-0.12	0.69	1.90	-2.04
	14/12/2017	0.08	-0.25	-2.36	-2.46	-0.18	-1.57	0.15	-0.13	-0.05	0.11	-2.44
	01/03/2018	-0.04	-0.44	-2.42	-2.49	-0.23	-1.54	0.11	-0.16	-0.39	-0.34	-2.26
DEMU	19/04/2016	0.03	-0.24	-2.89	-2.83	-0.77	-1.51	0.07	-0.19	-0.29	-0.41	-2.02
	18/07/2016	-0.12	-0.50	-2.83	-2.73	-0.82	-1.70	0.04	-0.22	-0.53	-0.72	-1.75
	04/10/2016	-0.08	-0.41	-2.89	-2.80	-0.88	-1.66	0.04	-0.22	-0.76	-1.09	-1.81
	10/01/2017	-0.03	-0.32	-2.86	-2.76	-0.90	-1.63	0.04	-0.22	-1.05	-1.67	-1.87
	21/03/2017	0.07	-0.15	-2.91	-2.87	-0.80	-1.49	0.12	-0.14	NC	NC	-2.10
	27/06/2017	0.11	-0.11	-2.85	-2.80	-0.84	-1.45	0.11	-0.15	-0.88	-1.60	-2.13
	03/10/2017	-0.06	-0.39	-2.87	-2.77	-0.83	-1.65	0.02	-0.24	-0.60	-0.79	-1.82
	14/12/2017	0.11	-0.06	-2.90	-2.84	-0.83	-1.60	0.09	-0.17	-0.36	-0.68	-2.10
	01/03/2018	-0.10	-0.50	-2.86	-2.81	-0.76	-1.66	0.11	-0.16	0.09	0.24	-1.90

*NC : Not Calculated

Highlights :

- Deep confined aquifers can be submitted to heavy and seasonal pressure variations
- Cyclic variations of sulfate and boron are observed over an extended area
- Concentration and isotope ratios of the two elements vary in accordance with each other
- Geochemical modelling based on water mixes fits the field observations
- Observations and simulations confirm water and mass fluxes within deep aquifers

Declaration of interests

☒ The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

☐ The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: