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# Geophysical Research Letters

## RESEARCH LETTER

10.1029/2019GL086239

### Key Points:

- First atmospheric distributions of acetic acid are obtained from IASI/Metop observations
- In emission regions, formic and acetic acids highly correlate with each other and with biogenic hydrocarbon emissions and biomass burning
- Residual smoldering combustion appears to be a major driver of the formic and acetic acids seasonality in Africa

### Supporting Information:

- Supporting Information S1

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## Spaceborne Measurements of Formic and Acetic Acids: A Global View of the Regional Sources

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**Abstract** Formic (HCOOH) and acetic acids (CH<sub>3</sub>COOH) are the most abundant carboxylic acids in the Earth's atmosphere and key compounds to aqueous-phase chemistry. Here we present the first distributions of CH<sub>3</sub>COOH retrieved from the 2007–2018 satellite observations of the nadir-looking infrared atmospheric sounding interferometer (IASI), using a neural network-based retrieval approach. A joint analysis with the IASI HCOOH product reveals that the two species exhibit similar distributions, seasonality, and atmospheric burden, pointing to major common sources. We show that their abundance is highly correlated to isoprene and monoterpenes emissions, as well as to biomass burning. Over Africa, evidence is provided that residual smoldering combustion might be a major driver of the HCOOH and CH<sub>3</sub>COOH seasonality. Earlier seasonal enhancement of HCOOH at Northern Hemisphere middle and high latitudes and late seasonal secondary peaks of CH<sub>3</sub>COOH in the tropics suggest that sources and production pathways specific to each species are also at play.

### 1. Introduction

Formic (HCOOH) and acetic acids (CH<sub>3</sub>COOH) are the primary atmospheric acids contributing to the cloud and rainwater acidity in remote environments (Galloway et al., 1982; Keene et al., 2015). They affect the oxidant concentrations and the pH-dependent reaction rates in cloud water (Chameides & Davis, 1983; Jacob, 1986; Vet et al., 2014). Due to the mitigation of the anthropogenic emissions of inorganic acid precursors like nitrogen (NO<sub>x</sub>) and sulfur oxides (SO<sub>x</sub>), the cloud and precipitation acidity is expected to be increasingly steered by organic acids (Galloway et al., 1982; Keene et al., 2015). Also present in the aerosol phase as formate and acetate ions, these organic acids serve as substrates for the formation of effective cloud condensation nuclei (Yu, 2000). Current knowledge indicates that the global sources of HCOOH and CH<sub>3</sub>COOH are dominated by atmospheric photochemical production via the sunlight-induced degradation of biogenic, anthropogenic, and biomass burning precursors (Khan et al., 2018; Millet et al., 2015; Müller et al., 2019; Paulot et al., 2011; Stavrakou et al., 2012). In particular, oxidation by the hydroxyl radical (OH) and ozonolysis of isoprene and several monoterpenes are thought to be the main formation pathways. The contribution from HCOOH and CH<sub>3</sub>COOH direct emissions is estimated not to exceed ~30% of the total sources (Müller et al., 2019; Paulot et al., 2011).

However, state-of-the-art global models incorporating the known emissions and secondary sources underpredict the measured atmospheric abundance of HCOOH and CH<sub>3</sub>COOH (Ito et al., 2007; Khan et al., 2018; Millet et al., 2015; Paulot et al., 2011; Stavrakou et al., 2012; von Kuhlmann et al., 2003). Since both acids are by-products of major biogenic hydrocarbons like isoprene and monoterpenes, these discrepancies might point to large gaps in our understanding of the atmospheric reactive carbon oxidation chains, and/or to the existence of one or several as-yet-unidentified source(s) of organic acids. In this framework, continuous and homogeneous global satellite observations may provide information on the nature of the missing sources and more constraints on their atmospheric budget. For instance, the first global distributions of HCOOH

(Pommier et al., 2016; Razavi et al., 2011) obtained from the nadir-looking Infrared Atmospheric Sounder Interferometer (IASI) satellite sensors (Clerbaux et al., 2009) and assimilated by the chemistry-transport model IMAGES, led to the identification of sizeable missing sources of HCOOH from the tropical and boreal forests (Stavrakou et al., 2012). The CH<sub>3</sub>COOH budget remains less constrained since, until now, CH<sub>3</sub>COOH has only been detected in IASI spectra recorded in exceptionally large fire events (Clarisse et al., 2011; Coheur et al., 2009). Its retrieval at large scale has so far not been attempted, due to the weak CH<sub>3</sub>COOH signature in the thermal infrared spectral range.

The development of the retrieval algorithm *Artificial Neural Network for IASI* (ANNI), version 3, recently allowed flexible retrievals for a suite of oxygenated volatile organic compounds (OVOCs) from the IASI observations (Franco et al., 2018, 2019), including new, robust daily global distributions of HCOOH columns. Relying on the ANNI framework, here we also perform the retrieval of CH<sub>3</sub>COOH vertical abundance throughout the observational time series of IASI/Metop-A (since 2007) and -B (since 2013). We present, for the first time, large scale daily maps of CH<sub>3</sub>COOH total columns from space, which we use to characterize the regional sources and seasonal variability of this compound. We also investigate the spatial correlations and ratios between the HCOOH and CH<sub>3</sub>COOH columns. Finally, regional time series of both species are studied alongside terrestrial biogenic emission fluxes and a biomass burning tracer.

## 2. Methods and Data

### 2.1. CH<sub>3</sub>COOH Detection in IASI Spectra

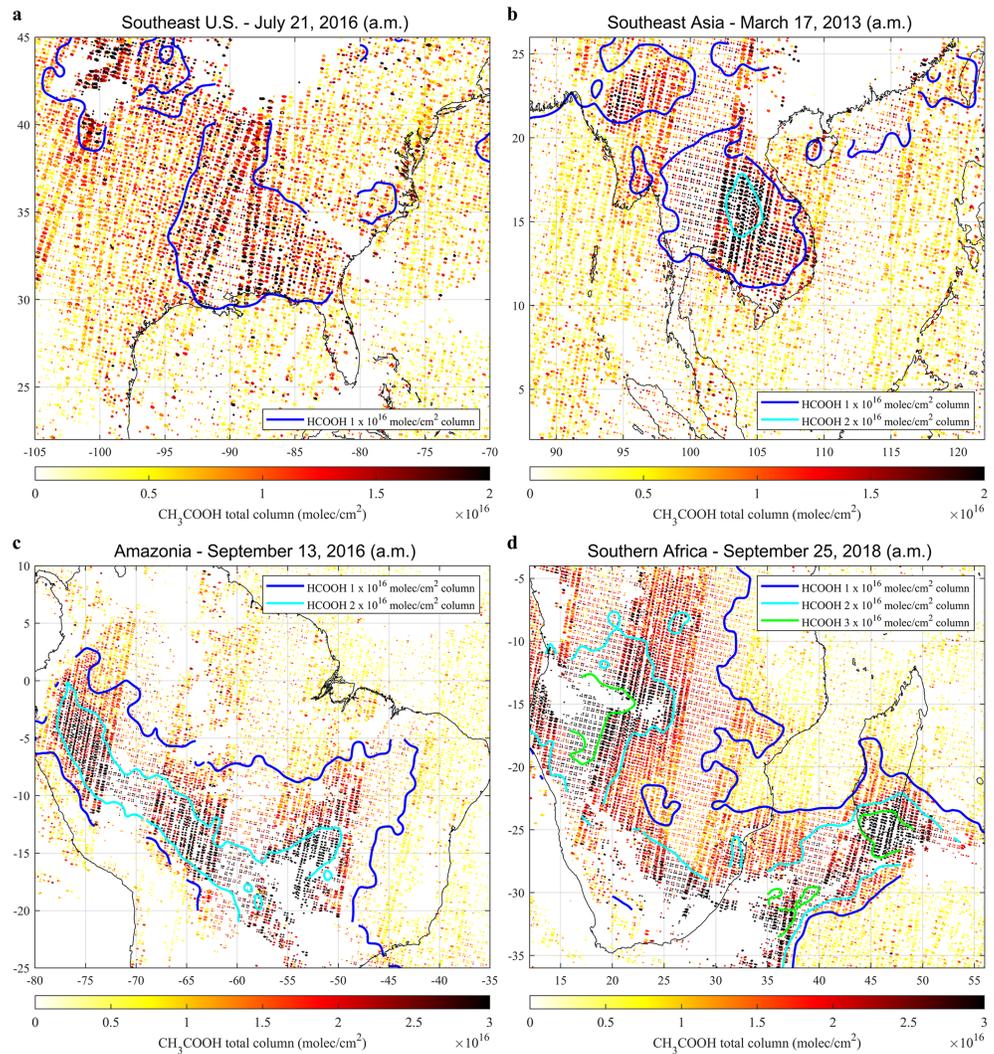
In the thermal infrared spectral range, the main absorption feature of CH<sub>3</sub>COOH away from water vapor interferences corresponds to its  $\nu_8$  band (CO stretching mode) in the 1,120–1,220 cm<sup>-1</sup> spectral window. This broad absorption band has already been identified in IASI spectra recorded during strong fire events in Eastern Siberia (Coheur et al., 2009) and the 2009 Australian bushfires (Clarisse et al., 2011). To demonstrate the detection of CH<sub>3</sub>COOH with IASI outside fire plumes, we selected over regional sources (section 3) IASI spectra with a high CH<sub>3</sub>COOH spectral index (see section 2.2), avoiding the fire plumes. This was done by filtering out measurements with elevated concentrations of biomass burning tracers, like carbon monoxide (CO) and ammonia (NH<sub>3</sub>). Examples are presented in supporting information Figure S10, which demonstrate the presence of the CH<sub>3</sub>COOH signature in IASI spectra. The spectral demonstration follows the same approach as Coheur et al. (2009), which consists in reconstructing the observed spectra successively with and without accounting for CH<sub>3</sub>COOH, the difference between the spectra visualizing the target gas signature (see Figure S10 for details).

### 2.2. Neural Network Retrieval

Retrieval of CH<sub>3</sub>COOH using traditional physical inversion is nonetheless very challenging to perform at the global scale, primarily due to the broadband, low-intensity CH<sub>3</sub>COOH absorption, and because of the presence in that spectral range of other broadband absorbers like PAN and acetone. Moreover, the difficulty of forward models to simulate accurately atmospheric infrared spectra over large spectral ranges constitutes an additional hindrance for the retrieval of such trace gases. Here, we adapt the ANNI retrieval method, initially developed for the retrieval of NH<sub>3</sub> (Van Damme et al., 2017; Whitburn et al., 2016) and dust (Clarisse et al., 2019) from IASI observations, and which has recently allowed global retrieval of HCOOH, methanol, PAN, and acetone (Franco et al., 2018, 2019). For the CH<sub>3</sub>COOH retrieval performed here, we follow closely the ANNI v3 procedure (Franco et al., 2018), for which an extended overview—including the steps and evaluation specific to CH<sub>3</sub>COOH—are provided in section 1 of the supporting information.

The ANNI retrieval method relies on two elements. The first is the so-called hyperspectral range index (HRI), which quantifies for each spectrum the magnitude of the target gas absorption in a given spectral range. Detection methods based on an HRI are typically much more sensitive than other detection methods, as the HRI can exploit the full information content of all spectral channels contained in a broad spectral range (Clarisse et al., 2013; Walker et al., 2011; Whitburn et al., 2016). Another advantage is that the HRI has a weak dependence on the forward model as it is only used for the calculation of a Jacobian (the covariance matrix can be estimated from real observed IASI spectra). In particular, using the HRI to quantify the absorption avoids relying on a full reconstruction of the spectrum and all the associated problems.

The other element of ANNI is a feedforward neural network (NN), which is used to convert the unitless HRI into a gas column abundance. While the HRI provides the ANNI method with its sensitivity to the



**Figure 1.** Regional  $\text{CH}_3\text{COOH}$  total column enhancements derived from individual days of IASI/Metop-A and -B observations (a.m. overpasses). The shape of each dot corresponds to the footprint on ground of an individual IASI measurement. The IASI/Metop-A and -B HCOOH column levels are displayed in contour lines.

target gas, the accuracy of the retrieval is brought by the NN that models as accurately as possible the functional relationship between the HRI and the target gas column. NNs are increasingly used for satellite data inversion as a robust, computationally efficient alternative to physical approaches (e.g., Aires et al., 2002; Di Noia et al., 2015; Müller et al., 2003; Piscini et al., 2014). They have the ability to accurately approximate complex systems at a fraction of the original computational cost. For this reason, they can replace atmospheric models and radiative transfer models (Krasnopolsky et al., 2005), which is the case here. Moreover, the NN computing efficiency permits the processing of the entire IASI/Metop-A and -B observational time series. Other advantages of the ANNI retrieval scheme are specifically discussed by Clarisse et al. (2019) and Whitburn et al. (2016).

Examples of regional  $\text{CH}_3\text{COOH}$  (and superimposed HCOOH) column distributions from individual days of IASI/Metop-A and -B observations are shown in Figure 1. The areas without data are due to measurements excluded because of clouds or poor retrieval performance. These distributions illustrate enhancements over Southeast United States and Southeast Asia (Figures 1a–1b) and transported plumes originating from Amazonia and Southern Africa (Figures 1c–1d), with column values of  $0.3\text{--}0.4 \times 10^{16}$  molec/cm<sup>2</sup> in remote regions and  $1.5\text{--}3 \times 10^{16}$  molec/cm<sup>2</sup> over source areas and in the middle of the plumes. Note the strong spatial correspondence between both organic acids. The corresponding  $\text{CH}_3\text{COOH}$  column uncertainties

typically range between 2 and  $8 \times 10^{15}$  molec/cm<sup>2</sup>, whereas the HCOOH uncertainties are a factor 2–3 lower (Figures S11–S12).

Independent column measurements of HCOOH are unfortunately sparse and do not even exist for CH<sub>3</sub>COOH. We nonetheless provide in section 2 of the supporting information a first comparison between the IASI products, HCOOH total columns derived from ground-based Fourier transform infrared (FTIR) observations made at Network for the Detection of Atmospheric Composition Change (NDACC) sites, and CH<sub>3</sub>COOH columns calculated from composite aircraft-model profiles. For HCOOH, these initial comparisons indicate a good agreement at most FTIR sites (with correlation coefficients usually higher than 0.6) and also biases at high-altitude stations or uncaptured local enhancements that will require further investigations. For CH<sub>3</sub>COOH, these comparisons are relatively inconclusive because of the lack and limited representativity of the comparison data, as well as the many uncertainties that are propagated throughout the comparison process.

### 2.3. Biogenic Emission Fluxes and Fire Data

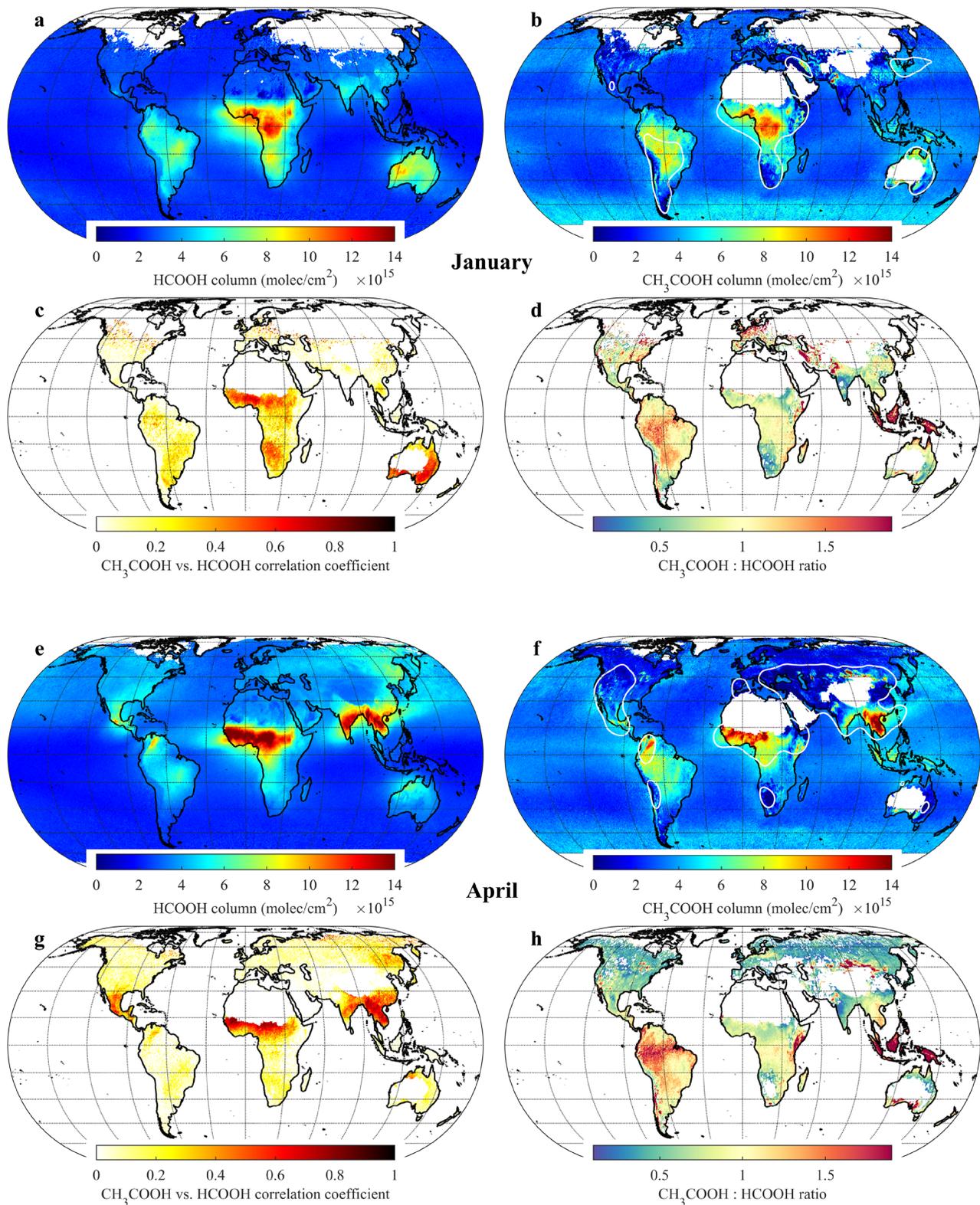
Hereafter, the correlations between biogenic hydrocarbon emissions, biomass burning, and the HCOOH and CH<sub>3</sub>COOH seasonal cycle are investigated. To represent the emissions of their main biogenic precursors, we use the fluxes of isoprene and monoterpenes—lumped together—calculated by the Model of Emissions of Gases and Aerosols from Nature (MEGAN; Guenther et al., 2012) coupled with the Model of HYdrocarbon emissions by the CANopy (MOHYCAN; Bauwens et al., 2018; Müller et al., 2008). To account for biomass burning, we rely on the detected fire radiative power (FRP) taken from the Level 2 Fire Products, Collection 6, of the MODerate resolution Imaging Spectroradiometer (MODIS) onboard the NASA Terra and Aqua satellites (Giglio et al., 2006). From this data set, we have computed monthly cumulative FRP by summing up all the individual fire contributions over specific areas. This quantity is a better indicator of the integrated biomass combustion than, for example, the fire count (which disregards the fire magnitude) or the average FRP (which ignores the number of fires).

## 3. Results

### 3.1. HCOOH and CH<sub>3</sub>COOH Distributions

Monthly global distributions of HCOOH and CH<sub>3</sub>COOH total columns, computed over the 2007–2018 IASI/Metop–A time series, are presented for January, April, July, and October in Figures 2–3 (all months are displayed in Figures S13–S14). Compared to the other OVOCs retrieved from IASI, the CH<sub>3</sub>COOH product is noisier and more sensitive to local spurious biases related to, for example, deserts, clouds, or water vapor. In addition, the CH<sub>3</sub>COOH signal is relatively close to the detection threshold, especially over remote areas ( $\sim 3\text{--}5 \times 10^{15}$  molec/cm<sup>2</sup> for individual observations). Therefore, the applied post-filter was chosen quite conservatively. For instance, most of the observations above Europe are removed during winter. However, even with this filter in place, small biases over land can still be present (e.g., in the Gobi desert) and the data over sea exhibit local spurious features due to persistent cloud coverage (e.g., around Antarctica). For these reasons also, the analysis focuses hereafter on regions with large CH<sub>3</sub>COOH enhancements. Typically, those are included in the areas delimited by the white contour lines in Figures 2–3 and S14, which correspond to the areas where the absolute uncertainty on a single-measurement basis (Figure S15) is below  $5 \times 10^{15}$  molec/cm<sup>2</sup> or where the relative uncertainty (Figure S16) does not exceed 100%. Provided that the measurements are not systematically biased, note that the actual data uncertainties are significantly lower owing to the numerous IASI measurements that are averaged in these CH<sub>3</sub>COOH distributions. Outside these areas, the confidence in the measurement is lower. It is worth mentioning that HCOOH is not affected to the same degree by these issues. Figures 2–3 also provide, for each month and over the 2007–2018 time period, the spatial correlation coefficients (*R*) between the daily HCOOH and CH<sub>3</sub>COOH column time series, and the CH<sub>3</sub>COOH:HCOOH column ratios.

The HCOOH and CH<sub>3</sub>COOH distributions exhibit strong continental enhancements which primarily take place in isoprene- and monoterpenes-dominated environments, namely, the tropical forests, Southeast Asia in spring, and Southeast United States and the boreal forests in summer (Figures 2–3). This is consistent with the major photochemical production pathways already identified for these acids that are the ozonolysis of VOCs with a double-bond moiety (Calvert & Stockwell, 1983; Neeb et al., 1997), the OH-oxidation of enol compounds (Lei et al., 2018; So et al., 2014) originating from acetaldehyde (Andrews et al., 2012) and isoprene by-products (e.g., Liu et al., 2018), and the reactions of the peroxyacetyl radical (CH<sub>3</sub>C(O)O<sub>2</sub>) with



**Figure 2.** (a–b) Means (on a  $0.5^\circ \times 0.5^\circ$  grid) of the HCOOH and CH<sub>3</sub>COOH total columns from the 2007–2018 IASI/Metop–A observations over January. The empty spaces at high latitudes are regions where no measurements passed the pre- and post-filtering. The CH<sub>3</sub>COOH data over deserts are not displayed. The white contour lines delimit the areas with high CH<sub>3</sub>COOH measurement confidence (see text). (c) Correlation coefficients between the daily  $0.5^\circ \times 0.5^\circ$  gridded HCOOH and CH<sub>3</sub>COOH total columns and (d) CH<sub>3</sub>COOH:HCOOH column ratios, over January throughout the 2007–2018 time period. (e–h) Same as (a–d), but for April.

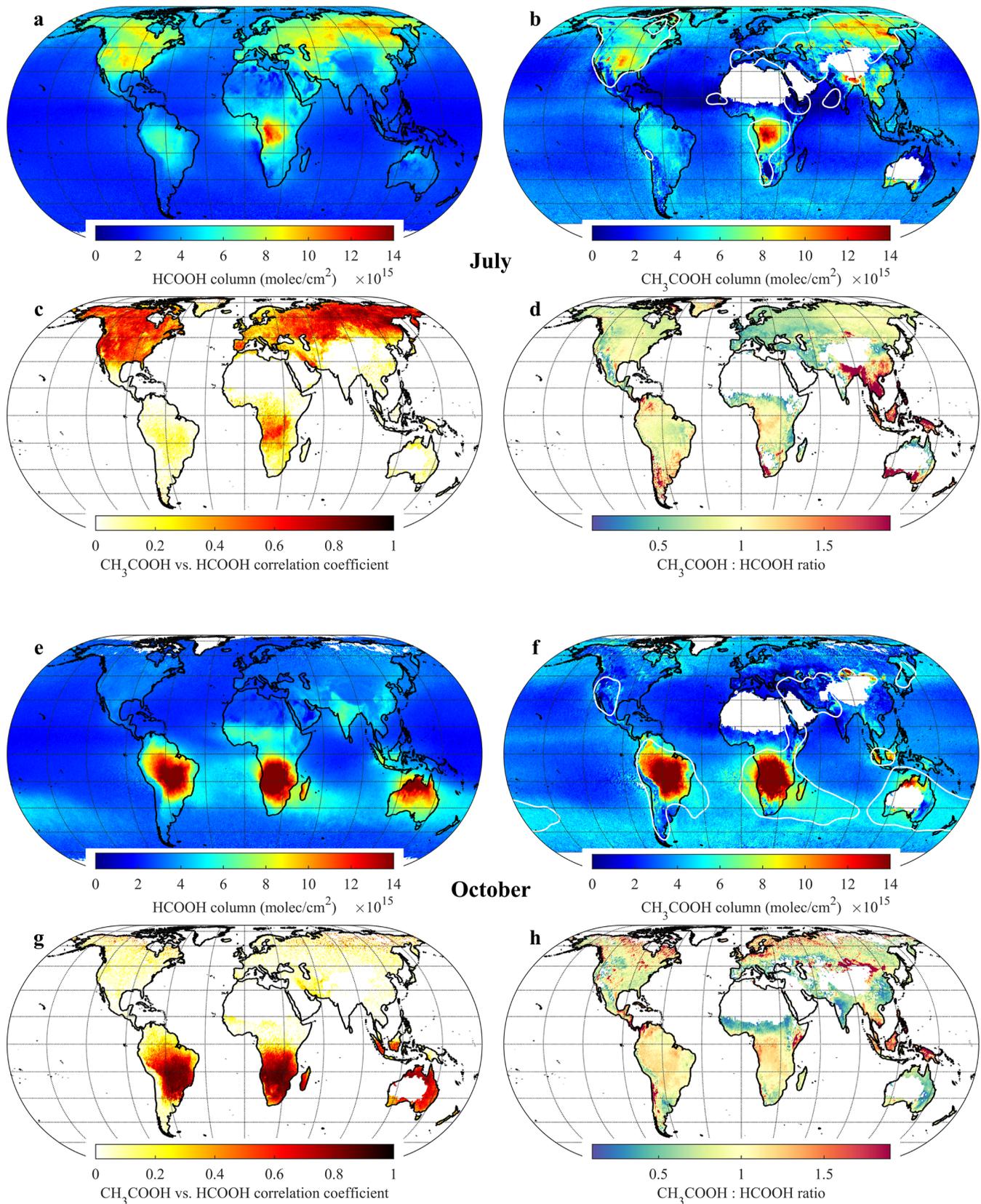


Figure 3. Same as Figure 2, but for July (a–d) and October (e–h).

the hydroperoxyl radical ( $\text{HO}_2$ ) and other peroxy radicals ( $\text{RO}_2$ ) in low  $\text{NO}_x$  concentration (Dillon & Crowley, 2008; Hasson et al., 2004; Moortgat et al., 1989). These enhancements are also due to biomass burning with occurrence, for example, in October over Southern Africa and Amazonia, in March–April in Southeast Asia and in summer in the boreal regions (Figures 2–3). Both organic acids and several of their precursors can indeed be directly emitted from fires, especially  $\text{CH}_3\text{COOH}$  which has 3 to 10 times higher biomass burning emission factors than  $\text{HCOOH}$  (Akagi et al., 2011; Andreae, 2019).

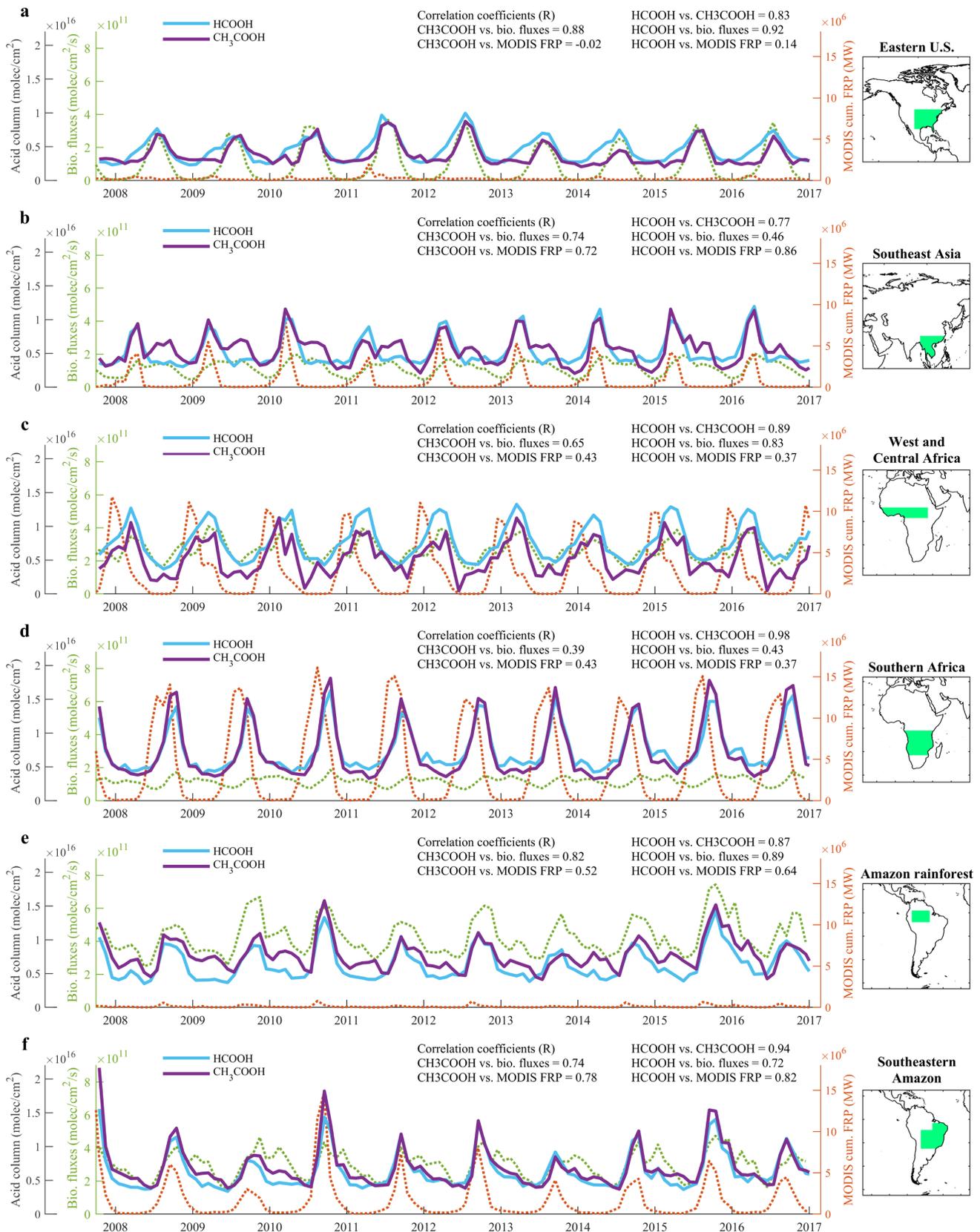
Overall, the IASI  $\text{CH}_3\text{COOH}$  measurements present spatial distributions and total columns that resemble those of  $\text{HCOOH}$ , with typical monthly averages of  $1\text{--}2 \times 10^{15}$  molec/ $\text{cm}^2$  in remote areas and  $>14 \times 10^{15}$  molec/ $\text{cm}^2$  over hot spots. Over source regions, where the  $\text{CH}_3\text{COOH}$  data are more reliable, the  $\text{CH}_3\text{COOH}:\text{HCOOH}$  ratios are close to 1 and the correlation coefficients higher than 0.5–0.6, for example, at Northern Hemisphere middle and high latitudes in July and over the biomass burning areas in October (Figure 3). This is in agreement with most field measurements that show a noticeable correlation between both species (e.g., Chebbi & Carlier, 1996; Keene & Galloway, 1988; Khare et al., 1999). Based on the emissions and photochemical production identified so far, the total  $\text{CH}_3\text{COOH}$  sources have been estimated at 73–134.3 Tg/year, against 32–61.5 Tg/year for  $\text{HCOOH}$  (Ito et al., 2007; Khan et al., 2018; Millet et al., 2015; Müller et al., 2019; Paulot et al., 2011; Stavrou et al., 2012; von Kuhlmann et al., 2003). However, the  $\text{CH}_3\text{COOH}$  degradation in the free troposphere through OH-induced oxidation is thought to proceed more rapidly than for  $\text{HCOOH}$  ( $\sim 10$  days against  $\sim 25$  days; Atkinson et al., 2006; Huang et al., 2009), which may partly explain the similar observed columns between the two species. Efficient removal from the atmospheric boundary layer via wet and dry deposition reduces their global lifetime to  $\sim 2.3$  and 3.2 days, respectively (Paulot et al., 2011); therefore, the abundance of both compounds drops off outside the source regions. Over the biomass burning areas, the ratios are similarly close to 1, although higher fire emission factors for  $\text{CH}_3\text{COOH}$  than for  $\text{HCOOH}$  are reported in the literature (Akagi et al., 2011; Andreae, 2019). While we cannot rule out potential retrieval biases, another explanation is a rapid secondary  $\text{HCOOH}$  formation, as suggested by space-based pyrogenic  $\text{HCOOH}$  enhancement ratios that are significantly higher than expected (Chaliyakunnel et al., 2016; Pommier et al., 2017), and which would not be present to the same degree for  $\text{CH}_3\text{COOH}$ .

IASI  $\text{HCOOH}$  and  $\text{CH}_3\text{COOH}$  distributions do not exhibit clear enhancements over polluted regions (e.g., East Asia and Europe). Indeed, only a few mechanisms of carboxylic acid production involving anthropogenic compounds have been identified so far, such as the  $\text{HCOOH}$  formation via the OH-oxidation of acetylene (Hatakeyama et al., 1986) and the ozonolysis of propene and higher alkenes leading to  $\text{CH}_3\text{COOH}$ . Moreover, whereas the  $\text{CH}_3\text{C}(\text{O})\text{O}_2 + \text{HO}_2/\text{RO}_2$  reactions yielding  $\text{CH}_3\text{COOH}$  dominate in clean and remote environments, in polluted conditions (high  $\text{NO}_x$  concentration)  $\text{CH}_3\text{C}(\text{O})\text{O}_2$  primarily reacts with  $\text{NO}_x$  to produce methyl peroxy radical ( $\text{CH}_3\text{O}_2$ ) and peroxyacyl nitrates. Although large direct  $\text{CH}_3\text{COOH}$  emissions have been reported locally from fossil fuel and biofuel burning (e.g., Kawamura et al., 1985; Liggio et al., 2017), altogether the primary and secondary anthropogenic sources are estimated not to contribute significantly to the  $\text{HCOOH}$  and  $\text{CH}_3\text{COOH}$  burden (Khan et al., 2018; Müller et al., 2019; Paulot et al., 2011), leaving the indirect biogenic source as the main known driver throughout the year.

Over all regions, we observe high background columns ( $1\text{--}2 \times 10^{15}$  molec/ $\text{cm}^2$ ) for both  $\text{HCOOH}$  and  $\text{CH}_3\text{COOH}$ . Although the background values of the IASI  $\text{HCOOH}$  measurements are generally trustworthy, as pointed out before the IASI  $\text{CH}_3\text{COOH}$  background data should be treated with caution. Interestingly, elevated concentrations in carboxylic acids over remote areas have also been reported by independent measurements (e.g., Chebbi & Carlier, 1996; Keene & Galloway, 1988; Khare et al., 1999; Mungall et al., 2018), but remain largely unexplained.

### 3.2. Regional Sources

Regional time series of monthly averaged IASI/Metop–A  $\text{HCOOH}$  and  $\text{CH}_3\text{COOH}$  columns are displayed in Figure 4, along with the cumulative MODIS FRP and MEGAN–MOHYCAN isoprene and monoterpenes fluxes. The gas columns show a clear peak in summer at Northern Hemisphere middle and high latitudes (Figures 4a and S17a), in spring over Southeast Asia, West and Central Africa, and Central America (Figures 4b–4c and S17b), and during the second half of the year in the Southern Hemisphere (Figures 4d–4f). The peak usually represents a factor 2–3 increase relative to the background levels. Overall, both species exhibit a similar seasonal cycle over all regions, which indicates the common origin of their



**Figure 4.** Time series of monthly mean HCOOH and CH<sub>3</sub>COOH total column (blue and purple solid lines, respectively), MEGAN–MOHYCAN isoprene and monoterpenes emission fluxes (green dotted line), and cumulative MODIS FRP (orange dotted line) over areas of interest.

sources and sinks. This is reflected in the large monthly correlation coefficients between the two species, with the lowest value ( $R = 0.77$ ) reached over Southeast Asia and the highest ( $R = 0.98$ ) over Southern Africa. The HCOOH and CH<sub>3</sub>COOH columns seem to be dominantly driven by the terrestrial biogenic hydrocarbon emissions over Eastern United States and the Amazon rainforest (Figures 4a and 4e), where the cumulated FRP remains weak. The Amazon rainforest area has been separated from the southeastern part based on the known biogenic sources of carboxylic acids, but fire plumes originating from the Amazon deforestation area also contribute to their seasonal enhancements. In contrast, despite a time lag (see below), the dependence on biomass burning seems to be the strongest over Southern Africa where the relatively constant biogenic fluxes cannot explain the HCOOH and CH<sub>3</sub>COOH seasonality (Figure 4d). In other regions, the data suggest a joint contribution from both terrestrial vegetation and fires. Indeed, whereas the biogenic emissions do not show strong interannual variability, a more intense yearly fire energy often translates into a stronger annual peak of organic acids, such as over Southeast Asia, the Amazon deforestation area, and the boreal forests (Figures 4b, 4f, and S17a).

Even if both organic acids highly correlate with each other, some differences are noticeable on a regional scale. For instance, HCOOH columns in Southeast United States (Figure 4a) starts increasing in late winter–early spring, that is,  $\sim 2$  months earlier than CH<sub>3</sub>COOH. The same feature is observed over the boreal forests (Figure S17a). Indeed, the CH<sub>3</sub>COOH:HCOOH ratio is significantly lower than 1 in April at Northern Hemisphere middle and high latitudes (Figure 2h). Another major difference is seen over the low latitude regions (except over Africa), where the post-peak return of the CH<sub>3</sub>COOH columns to the background values is slower than for HCOOH. This slow decrease furthermore exhibits a secondary peak over Southeast Asia, the Amazon rainforest, and Central America (Figures 4b, 4e, and S17b), which does not exist for HCOOH. Owing to the limited sensitivity of IASI to the CH<sub>3</sub>COOH detection outside the enhancement periods, we cannot at this stage rule out the possibility that these differences in the time evolution are artifacts. However, should independent measurements corroborate these observations, this would imply that one or several direct source(s) and/or chemical mechanism(s) are seasonally active which do not lead to the production of the other species.

Over Africa (Figures 4c–4d), the HCOOH and CH<sub>3</sub>COOH seasonality seems to be closely related to the fire seasonal cycle (Figures 2–3), which typically starts with the savanna fires in the northern sub-Saharan regions in November–December, proceeds southward, and finishes in Southern Africa in October–November the next year. The peak of HCOOH and CH<sub>3</sub>COOH columns rather occurs in the late regional fire season, that is, 1–2 month(s) later than the peak of MODIS FRP taking place usually in the early dry season. This time lag has been reported in the literature for other fire tracers, for example, with IASI CO and NH<sub>3</sub> column measurements (Whitburn et al., 2015). Recently, this has been ascribed to the transition from flaming- to smoldering-dominated combustion during the course of a fire season (Zheng et al., 2018); the late season shows more precipitation and declining FRP, which decreases the combustion efficiency and increases the emission factors of incompletely oxidized products, like CO, organic acids, and several of their precursors. A shift from grassland to woodland fires lowering the combustion efficiency has also been evoked (van der Werf et al., 2006). It is worth mentioning that the observed enhancements match also with the biogenic emissions over west and central Africa (Figure 4c). Interestingly, no clear time lag is observed over other typical biomass burning areas, such as Southeast Asia and Southeastern Amazonia (Figures 4b and 4f). The same time lag over Africa is observed for CO with bottom-up fire emission inventories that rely on satellite-derived FRP and burned area (e.g., Zheng et al., 2018). Since these inventories implement seasonally static emission factors and do not account for the flaming to smoldering transition, this suggests an underestimation of the HCOOH and CH<sub>3</sub>COOH emissions and/or precursors during the late fire season.

#### 4. Conclusions

We have produced and characterized the first distributions of CH<sub>3</sub>COOH total columns obtained from IASI observations. The IASI CH<sub>3</sub>COOH product complements the suite of ANNI OVOCs retrieved globally. Studying the IASI HCOOH and CH<sub>3</sub>COOH columns together indicates a strong dependence of both species to the major hydrocarbon emissions from the terrestrial vegetation over almost all the regions where organic acid enhancements are detected. Biomass burning emissions of HCOOH and CH<sub>3</sub>COOH and of their precursors contribute significantly to these hot spots and to the interannual variability in the tropics, especially over

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Africa. Conversely, the absence of clear enhancements over major polluted regions sustains the hypothesis that anthropogenic emissions contribute little to the atmospheric burden of carboxylic acids.

Although HCOOH and CH<sub>3</sub>COOH share common sources, the IASI observations suggest the existence of emissions and/or routes leading specifically to the production of each species, which should be verified by independent data. Differentiated measurements of biogenic tracers would also be required to identify the HCOOH and CH<sub>3</sub>COOH biogenic precursors and the mechanisms that are at play. Also, a more advanced proxy of combustion efficiency would help to better represent the transition between dominant flaming combustion and residual smoldering, which is most likely the cause of the 1–2 month time lag over Africa between the peaks of MODIS fire tracer and of IASI organic acid columns.

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