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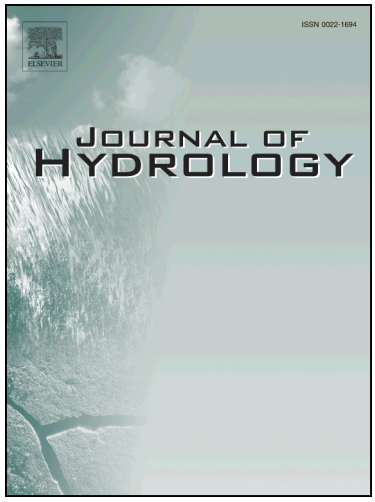
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Recharge processes and vertical transfer investigated through long-term monitoring of dissolved gases in shallow groundwater

V. de Montety[1]*, L. Aquilina[2], T. Labasque[2], E. Chatton[2], O. Fovet[3], L. Ruiz[3], E. Fourré[4], J.R. de Dreuzy[2]

[1] HydroSciences, UMR 5569, Université de Montpellier, Place E. Bataillon, 34095 Montpellier Cedex 5, France
[2] OSUR-Géosciences Rennes, Université Rennes1-CNRS, 35042 Rennes, France
[3] INRA, Agrocampus Ouest, UMR1069 Sol Agro et hydrosystème Spatialisation, 35000 Rennes, France
[4] LSCE, UMR8212, IPSL/CEA – CNRS – UVSQ, Saclay, France

*Corresponding Author:
Email: veronique.de-montety@univ-montp2.fr
Phone: +33 (2) 2323-5469

Abstract

We investigated temporal variations and vertical evolution of dissolved gaseous tracers (CFC-11, CFC-12, SF_6, and noble gases), as well as $^3H/^3He$ ratio to determine groundwater recharge processes of a shallow unconfined, hard-rock aquifer in an agricultural catchment. We sampled dissolved gas concentration at 4 locations along the hillslope of a small experimental watershed, over 6 hydrological years, between 2 and 6 times per years, for a total of 20 field campaigns. We collected groundwater samples in the fluctuation zone and the permanently saturated zone using piezometers from 5 to 20m deep. The purpose of this work is *i*) to assess the benefits of using gaseous tracers like CFCs and SF6 to study very young groundwater with flows suspected to be heterogeneous and variable in time, *ii*) to characterize the processes that control dissolved gas concentrations in groundwater during the recharge of the aquifer, and *iii*) to understand the evolution of recharge flow processes by repeated measurement campaigns, taking advantage of a long monitoring in a site devoted to recharge processes investigation.

Gas tracer profiles are compared at different location of the catchment and for different hydrologic conditions. In addition, we compare results from CFCs and $^3H/^3He$ analysis to define the flow model that best explains tracer concentrations. Then we discuss the influence of recharge events on tracer

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concentrations and residence time and propose a temporal evolution of residence times for the unsaturated zone and the permanently saturated zone. These results are used to gain a better understanding of the conceptual model of the catchment and flow processes especially during recharge events.

Key words: CFCs, SF_6, noble gases, residence time, groundwater recharge processes, flow model, agricultural catchment, vadose zone

1. Introduction

As nitrates have become a major environmental threat in agricultural areas, numerous works evaluate mechanisms controlling NO_3^- exportation from catchment and emphasize the key role of the shallow groundwater (Böhlke and Denver, 1995; Creed and Band, 1998; Martin et al., 2006; Molenat et al., 2002). In particular, these studies show that catchment can present an important time lag in response to variation of the NO_3^- input signal. Although physical characteristics of shallow groundwater are pointed out to explain this delay, mechanisms involved are still debated. A better knowledge of this time lag, through the evaluation of the residence time of groundwater is thus crucial for managing groundwater in small agricultural catchments.

Recent results obtained using natural solutes and/or their isotopic composition as tracers, pointed out that small catchments show variable residence times that can surprisingly exceed several years (Hrachowitz et al., 2013; Kendall and McDonnell, 1998; Kirchner et al., 2001; Ruiz et al., 2002a). These methods were based on opposite trends observed between the nitrate concentrations in drainage water and in stream water (Ruiz et al., 2002a) or on spectral analysis of the input-output signal of chemical compound (Kirchner et al., 2001; Molénat et al., 2000; Neal and Kirchner, 2000) that require a high frequency monitoring of rain inputs for a long period which is rarely available.

Another approach to estimate response time to changes in agricultural practices consists in numerical modelling. Again, transit times greater than one year and up to 14 years in headwater catchments have

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been derived from mechanistic groundwater modelling (Basu et al., 2012; Martin et al., 2006; Molénat and Gascuel-Odoux, 2002) or lumped and parsimonious model and time series of nitrate concentration in streams (Fovet et al., 2015; Ruiz et al., 2002b). All these approaches use the approximation of Dupuit-Forsheimer which reproduces the water dynamic of the catchment, but assume a perfect and instantaneous vertical mixing of groundwater neglecting vertical flows. In addition, in numerous models, individual storm events are neglected and only base flow is considered (Fovet et al., 2015; Ruiz et al., 2002b). Although these models reproduce seasonal and inter-annual nitrate variations, processes inducing temporal concentration variations at shallow depth are still poorly understood, especially after storm events.

An interesting alternative to obtain additional information on water dynamics is to analyze anthropogenic atmospheric gases. Industrial production of CFCs and SF_6 started in the 1940s and 1960s respectively which makes these dating tracers suitable for the study of post-war diffuse pollution in agricultural catchments (IAEA, 2006). However, following Montreal protocol in 1987, stagnation and then decrease of atmospheric CFCs concentrations over the past two decades have induced higher uncertainties in dating of young shallow groundwater since two different recharge dates can be derived from a single gas concentration (rising or falling part of the atmospheric curve (CMDL/NOAA)). Therefore, the combined use of other dating tracers like SF_6 (whose input atmospheric function is always rising) and $^3H/^3He$ (a radioactive parent/daughter clock method which does not require the knowledge of the ^3H input function) allows reducing uncertainties (Darling et al., 2012; Newman et al., 2010; Suckow, 2014). Another advantage of the anthropogenic gas tracers is to help conceptualizing flow and mixing processes of the studied groundwater system (Darling et al., 2012; Suckow, 2014). It is common to use lumped parameter models which fit the data to a conceptual model of defined flow, among which the most commonly used are piston (PFM), binary mixing (BMM) or exponential mixing (EM) (Maloszewski and Zuber, 1982). While each different tracer will give an apparent tracer age, stating the model used to interpret tracers concentrations will allow defining a Mean Residence Time (MRT) corresponding to the weighted average of the age distribution (Suckow, 2014).

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Numerous authors used CFCs and SF_6 to obtain a better understanding of pollutant-transfer processes and to predict the most likely evolution of groundwater quality with respect to diffuse agricultural pollution (Böhlke, 2002; Gooddy et al., 2006; Katz et al., 2001; Koh et al., 2006). Gooddy et al. (2006) used the comparison of the spatial distribution of different tracers and lumped parameter models to discriminate groundwater flow regimes in different parts of the aquifer and thus improved their conceptual model. Kolbe et al. (2016) coupled groundwater modelling with CFC age dating to classify local groundwater circulation in an unconfined aquifer. Recent studies in hard-rock aquifers in Brittany provided a potential regional distribution of residence times as well as a reconstruction of NO_3^- concentration in recharge water (Aquilina et al., 2012; Ayraud et al., 2008).

Temporal variability of residence time in the first meters of the aquifer usually receives less attention though being essential to further assess recharge processes in the critical zone. For instance, few studies focus on vertical variation of CFCs in shallow groundwater although it is a key parameter to constraint solute transfer from soil to the water table and within the aquifer (Cook et al., 1996, 1995; Le Gal La Salle et al., 2012). Moreover, no study to date investigates the temporal variability of dissolved CFCs and SF_6 concentrations at a single location over several hydrologic cycles nor how robust is the conceptual model over time. Such information can be used to evaluate how residence time changes with varying recharge conditions.

In this context, the scope of this paper is *i*) to assess the benefits of using gaseous tracers like CFCs and SF6 to study very young groundwater with flows suspected to be heterogeneous and variable in time, *ii*) to characterize the processes that control dissolved gas concentrations in groundwater during the recharge of the aquifer, and *iii*) to understand the evolution of recharge flow processes by repeated measurement campaigns, taking advantage of a long monitoring in a site devoted to recharge processes investigation (Legout et al., 2005; Rouxel et al., 2011).

To address these issues, we focus on a shallow unconfined aquifer in a small experimental catchment in Brittany (West of France) where flow processes and nitrate dynamic have been extensively studied giving a strong knowledge on water quality (Martin et al., 2006, 2004; Rouxel et al., 2011; Ruiz et al., 2002b).

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We investigate dissolved CFCs and SF_6 gases concentrations during 6 hydrological cycles at different depths ranging from 4 to 20 m below ground level. Gas tracer profiles are compared at different locations of the catchment and for different hydrologic conditions. We compare CFCs and $^3H/^3He$ apparent ages and discuss flow model that best explains tracer concentrations and origin of apparent age discrepancies between age dating tracers. Then we discuss the influence of recharge events on tracer concentrations and residence time and propose a temporal evolution of residence times for the unsaturated zone and the permanent groundwater of the catchment. These results are used to gain a better understanding of the conceptual model of the catchment and flow processes especially during recharge events.

The originality of this work is thus *i*) to use a detailed analysis of long-term time-series of dissolved gas tracers (CFC-11, CFC-12, SF_6 as well as recharge temperature and excess air deduced from noble gases concentrations); *ii*) to investigate their temporal evolution in vertical profiles ranging from 0 to 20 m below ground level.

2. Methods

2.1. Site description

Kerrien site is a small experimental catchment (0.095km^2) located closed to the sea in an intensive agricultural area south-west of Brittany, France (Figure 1). This site is part of the French observatory network RBV (http://rnbv.ipgp.fr/) and belongs to the Environmental Research Observatory AgrHys (http://www6.inra.fr/ore_agrhys_eng/) devoted to studies of response times of hydro-chemical fluxes under changing agriculture.

The study site is representative of hard rock catchments. It has been previously described by Martin et al. (2004). The fissured and fractured granitic bedrock is overlain by a regolith of an estimated average thickness of 20m (Legchenko et al., 2004 in Martin et al., 2004). The regolith constitutes the main

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unconfined, shallow aquifer, with a hydraulic conductivity ranging from 9.10^{-6} to 5.10^{-4}m.s^{-1} and a total porosity of 0.4% (Rouxel et al., 2012). The hillslope of the catchment shows a topographic gradient varying from 14% upslope to 5% downslope. Groundwater table roughly follows the topographic slope. It comes up to the ground level downslope while upslope, a zone of about 2m thickness remains permanently unsaturated. The surface runoff is negligible on the hillslopes and the shallow groundwater provides most of the stream flow with a base flow index about 90% (Ruiz et al., 2002b). Groundwater table level displays large seasonal fluctuations ranging from less than one meter downslope to 4 to 6 meters upslope.

The climate is oceanic. Mean annual rainfall for the period 2003 - 2010 was 1033 ± 186mm, slightly lower than the mean annual rainfall of 1185mm over the decade 1992 - 2002 (Molenat et al., 2008). The interannual variability is large with rainfall in 2005 (616mm) being half of rainfall in 2010 (1220mm). Mean annual PET is less variable with a mean annual value of 704 ± 19mm. The average maximum and minimum monthly rainfall occurs in November and September (143mm and 47mm respectively). Recharge period extends from November to March but it can vary slightly depending on the yearly rainfall pattern.

Groundwater chemical composition has been extensively studied since the early 1990s in this site (Martin et al., 2004; Martin et al., 2006; Molenat et al., 2008; Rouxel et al., 2011, and references therein). Groundwater has been affected by diffuse pollution due to intensive agriculture and breeding over the last 40 years (Ruiz et al., 2002b). The groundwater display high nitrate concentrations upslope and below 2m downslope (Molenat et al., 2008). Nitrate fluxes and seasonal variability have been analyzed in previous works (Martin et al., 2004; Martin et al., 2006; Rouxel et al., 2011; Ruiz et al., 2002b).

The site was instrumented since the early 1990s. In December 2004, 42 nested piezometers from 3 to 15m depth were drilled at three different positions along the slope. They were organized in two transects along the hillslope, each containing 3 nests of 7 piezometers (Figure 1). The shallow wells (3 and 4m) were screened on the last 50 cm only and the other ones (5, 6, 8, 10 and 15m) were screened on the last meter,

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allowing precise sampling at the different depths (Rouxel et al., 2011). This study focuses on the first transect H, J, L. After the installation of the piezometers, the catchment received low input of mineral fertilizer or manure (Rouxel et al., 2011). In addition, we sampled a 20m deep piezometer (F), screened between 19m and 20m, and drilled in 2001 upslope near the catchment divide (Martin et al., 2004).

2.2. Sampling and chemical analyses

Field Sampling

Groundwater samples were collected from the piezometers at the 3 nests H, J, L at depth ranging from 5 to 15m and from the piezometer F at 20m depth (Table 1). Twenty sampling surveys have been realized during the monitoring period covering 6 hydrological cycles. More than 170 samples were collected for CFCs analyses (2005 to 2010) and 110 samples for SF_6 (2006 to 2010). Noble gases have been sampled since 2007. The number of samples per piezometer varied from 3-5 samples for shallow piezometers (5m) to up to 16 samples in the permanent groundwater (10-15m).

Groundwater samples were collected with a submersible MP1 Grundfos pump connected to a nylon tubing after stabilization of their physico-chemical parameters (temperature, specific conductivity, pH and ORP). CFCs and SF_6 samples were collected without atmospheric contact in glass ampoules of 20ml (CFCs) and 500ml (SF_6) closed with two PTFE three-way valves for samples before March 2007 and in stainless-steel ampoules of 40ml (CFCs) and 300ml (SF_6) closed with two stainless-steel three-way valves since then. Noble gases were sampled in 500ml glass bottles after purging three times the bottle volume under water in a bucket. Bottles were closed with rubber caps and sealed with metal ring. All sampling ampoules were rinsed 3 times before collecting water. Physico-chemical parameters were simultaneously measured and major elements collected for each sampling period. Samples were filtered in the field (0.45µm) and kept refrigerated prior to analyses. In addition, for four sampling campaigns (June07, Dec08, Apr and Oct09) specific sampling was dedicated to $^3H/^3He$ analyses, representing 39 samples. Samples for ^3H analyses were collected in 500ml Pyrex bottles, baked at 75°C and pre-filled with argon.

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Samples for He analyses were collected using standard refrigeration grade 3/8" copper tubes sealed by metal clamps at both ends.

Water table levels were recorded at 15min intervals in the deepest piezometers (H15, J15, and L15). Precipitation, and the parameters required to calculate potential evapotranspiration by the Penman formula (mean temperatures, mean relative humidity, global radiation, wind velocity) were recorded at hourly frequency 700m away from the study site.

Lab analyses

Anthropogenic and noble gases analyses were realized in Geosciences Rennes laboratory as detailed in Labasque et al. (2008). CFCs and SF_6 were extracted following the purge and trap method and quantified by gas chromatography with electron capture detection. The analyses are within a precision of 3% for CFC-11 and CFC-12 and 5% for CFC-113 and SF_6. Noble gases (Ne, Ar) and N_2 were extracted by head-space and measured by gaz chromatography with a catharometer detector (µGC 3000 - SRA) with a precision of 5%. Major anions (Cl^-, SO_4^-, NO_3^-) were analyzed with an automated Dionex DX-100 Ion Chromatograph with a precision better than 5%. Tritium and noble gas isotopes (^3He, ^4He and ^{20}Ne) were analyzed at LSCE (Saclay, France) with a MAP-215-50 mass spectrometer using routine procedures (Jean-Baptiste et al., 2010, 1992). ^4He and ^{20}Ne isotopes were measured to assess excess air and radiogenic contribution and derive the tritiogenic ^3He component. Dissolved He and Ne were first extracted under vacuum into sealed glass tubes. The measurements were calibrated with an air standard. The water for ^3H determination was degassed and stored to allow for ^3He ingrowth before mass spectrometry measurement. Mean uncertainty is about 5% for ^3H, and between 1% and 1.5% for He and Ne isotopes.

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2.3. Data analyses

Uncertainty of apparent CFCs and SF6 age determination relies first on the calculation of atmospheric mixing ratio (pptv) in recharge water. This calculation requires the groundwater recharge elevation and the recharge temperature. The studied catchment can be assumed as a simple homogeneous system with an average elevation of 30m above sea level and low topographic variation within the catchment area (< 10 m). Influence of recharge elevation is reported to be small for elevation variation below 1000m (IAEA, 2006), the recharge elevation has thus a minor impact in CFCs determination in this study. The recharge temperature can be assumed to be close to the mean annual air temperature ± 1°C (Cook et al., 2006; Herzberg and Mazor, 1979). During the monitoring period, the mean annual air temperature is 11.9±0.4°C, in good accordance with previous work (11.4°C; Ruiz et al., 2002b), with mean monthly temperature ranging from 6.4°C in February to 16.7°C in July.

Ne and Ar concentrations were also used to estimate Noble Gaz recharge Temperature (NGT) as well as excess air. Two methods were used: (i) Ar/Ne chart (Heaton and Vogel, 1981) (ii) an inversion model (Aeschbach-Hertig et al., 2000; Stute et al., 1995) using, Ne, Ar and N2 to compute the recharge parameters including recharge temperature, excess air and fractionation using either a closed equilibrium or a partial re-equilibration model (Chatton et al., 2016). A comparison of the three computations showed only a slight difference of recharge temperature (0.6°C). Therefore, the closed equilibrium model has then been used.

At the water table (6m), water temperature varies between 11±1°C in March to 13±1°C in the period October-December. Mean NGT is about 10°C with variations between 7.5°C and 17°C with most of the data comprise between 7°C and 12.5°C. There is a strong variation of NGT (±3°C) compared to water temperature in wells with numerous data below the lower measured water temperature. When considering impact of recharge temperature on calculated CFCs ratios, it appears that the use of low temperature value (7°C) increases solubility and then decreases CFCs ratios for the same concentration in water. As a consequence, permanent groundwater shows older apparent tracer ages, (162pptv -26yrs at 7°C compare

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to 214pptv -20yrs at 11.9°C for CFC-11) and ratios in shallow water become too low for estimating age on the recent part of the atmospheric curve. Finally, we do not have NGT for all the dataset since noble gas measurement started only in 2007. For all these reasons and to not introduce a bias in the CFCs ratios calculation at the different sampling date, mean annual air temperature has been used for CFCs and SF6 ratios calculation. NGT temperature estimated with only two noble gases (Ne and Ar) gives too large uncertainties. Kr and Xe determination should be developed. N2 utilization in this agricultural watershed should also introduce uncertainties on NGT. The sensitivity of apparent age determination to a variation of 1°C of recharge temperature is small. The sensitivity for CFC-11 is less than 2 years.°C^{-1} for recharge before 1994 and 5 years.°C^{-1} in average for recharge after 1994. For CFC-12, only recharge prior to 2002 were determined with apparent ages precisions about 2 years.°C^{-1}.

SF_6 data were corrected for the influence of excess air using Ne concentrations according to the methodology proposed by Aeschbach-Hertig (2004). For sample collected before the first noble gases sampling, a mean value of excess air was calculated for each piezometer using samplings from 2007 to 2010. The influence of excess air was found to be negligible for CFCs and has thus been neglected.

The calculated mixing ratios were then compared to atmospheric mixing curves to estimate the apparent age of water. Considering the rural location of the study site, no local contaminations were expected (Ayraud et al., 2008). The CMDL/NOAA values were thus used as input function for CFCs and SF_6. CFC-11 concentrations and to a lesser extent CFC-12 concentrations fall close to the maximum atmospheric air concentration due to the relatively short residence time of water in this shallow aquifer. Consequently, for samples with concentration ranging from 240pptv to 272pptv for CFC-11 and from 530pptv to 546pptv for CFC-12, 2 apparent recharge dates could be attributed (before or after the atmospheric maximum value). If all samples had been attributed to the same portion of the atmospheric curve, the same range of apparent age would have been attributed to shallow (4 to 6m) and deep samples. Moreover, when considering apparent recharge date before the maximum atmospheric concentration for samples from 6 to 10m, a general decrease of apparent age with depth was observed. These two results

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were found unrealistic and were in strong disagreement with observed Cl^- evolution and prior knowledge on the hydrogeological context. To decide which portion of the atmospheric curve should be used, we used the vertical evolution of $^3H/^3He$ age. $^3H/^3He$ analysis show low age for samples at 6m (<2 years) and a general increase of age with depth (~ 0.75 yrs/m). Consequently, recharge dates were determined using the portion of the atmospheric curve before the maximum for deep groundwater (15-20m) and the portion of the atmospheric curve after the maximum for the other depth (10, 8, 6, 5m).

The $^3H/^3He$ age was computed using usual methodology (Schlosser et al., 1989). The $^3H/^3He$ age τ is defined as $\tau = \lambda^{-1}Ln(1+{}^3He_{tri}/{}^3H)$, where λ is the decay constant for tritium; ^3H is the measured tritium concentration expressed in Tritium Units (TU) and $^3He_{tri}$ is the fraction of the total ^3He produced by ^3H decay. $^3He_{tri}$ is the difference between the measured ^3He concentration and the sum of ^3H concentrations from atmospheric ($^3He_{atm}$) and terrigenic ($^3He_{ter}$) origins. The atmospheric component reflects both the gas dissolved at the solubility equilibrium and an excess air component as ubiquitously observed in groundwater (Kipfer et al., 2002 and references therein). $^3He_{atm}$ and $^3He_{ter}$ can be deduced from the measured concentrations of He and Ne, $^4He_{meas}$ and $^{20}Ne_{meas}$, using the following set of equations:

- $^4He_{atm} = {}^{20}Ne_{meas} \times \alpha$, where α is the solubility of $^4He/^{20}Ne$ at the recharge temperature (12±1°C),
- $^3He_{atm} = {}^4He_{atm} \times R_a \times 0.985$, with the atmospheric $^3He/^4He$ ratio $R_a = 1.38 \times 10^{-6}$,
- $^3He_{ter} = ({}^4He_{meas} - {}^4He_{atm}) \times R_{ter}$ where R_{ter} stands for the isotopic ratio of radiogenic helium production in rocks, set in this work at the typical value of $2.0 \pm 0.5 \times 10^{-8}$.

The overall uncertainty in the apparent $^3H/^3He$ ages is about 3 years in the case of this study.

3. Results

3.1. Hydrodynamic and chemical characterization of the study period

Hydrological conditions

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Contrasted hydrological conditions were observed leading to differentiate two hydrological periods and one particular recharge event on the whole data set. The two first cycles (2005 and 2006) have precipitations below the mean inter-annual value (Figure 2A) and correspond to a dry period. During this period, daily water table variations are low even during the recharge period. From 2007 to 2010, precipitations are higher or equal to the inter-annual value and characterize a humid period. Water table shows high daily variations in the early stage of the recharge periods (Figure 2C). In particular, the hydrological year 2007 is characterized by intense precipitation inducing a very sharp increase of water table at the beginning of the recharge period (fall 2006). However, according to Ruiz et al. (2002b), on average, the fast flow associated with such storm events represents only 10% of the total annual flux.

Groundwater level shows high amplitude of seasonal variations. Upslope, close to the watershed divide, (nest H; Figure 2B), water table depth varies from 5.5 - 6m to 2 - 3m during the hydrologic cycle. The permanently unsaturated zone is thus about 3m thick. The aquifer remains permanently saturated up to 6m below ground level. Water table level is slightly deeper downslope (nest J and L; Figure 2B) but fluctuations observed upslope (nest H) and downslope (L) show a similar pattern. Midslope nest (J) shows a slightly different pattern with sharper fluctuations and the deepest water table level at the end of the recession period reflecting lower permeability in this area.

Influence of recharge and recession periods on groundwater chemistry

As previously reported Cl^- and NO_3^- show important annual variations (Figure 2C, NO_3^- not shown) which contrast with the relative stability of the inter-annual values (Legout et al., 2007; Ruiz et al., 2002b). During the early stage of recharge periods shallow piezometers show a sharp decrease of Cl^- concentrations particularly visible for the recharge event of fall 2006. Such Cl^- drops indicate rapid circulations through the vadose zone inducing important dilution effect down to 6m depth. Such events can be characterized as “by-pass” events. After 2007, the lack of data for these piezometers does not allow the observation of a similar evolution. However, a similar sharp decrease has been recorded on conductivity (not shown), confirming the dilution effect for early stages of recharge period. Then, after

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2006, Cl^- variations follow water table fluctuations: during second stage of recharge period, Cl^- concentration increases when water level increases and reaches the highest Cl^- concentration at the end of the recharge periods (Figure 2C). This variation has been previously suggested to be related to soil and regolith matrix contribution (Legout et al., 2007).

During recession period, Cl^- concentration decreases following water table decreases and reaches the lowest Cl^- concentration of the cycle. Cl^- variations are observed both in unsaturated and saturated zone down to 15m but the variations are larger for shallow piezometers. In contrast, the deepest piezometer (F; 20m depth), which presents the lowest Cl^- concentration, shows almost no annual variation. This evolution shows that the deep part of the saturated zone is not affected by preferential vertical flow path and thus, its chemical composition can be considered as representative of the groundwater end-member to interpret the highly variable chemical composition of shallow piezometers. As the dynamic of Cl^- concentration is particularly visible during the 2006-2007 cycle, we will therefore focus on this cycle.

3.3. Recharge temperature and excess air in groundwater

Figure 3 shows the recharge temperature and excess air derived from the Ne, Ar and N_2 concentrations. Recharge temperatures range from 7.5 to 17°C with most of the data between 7.5 and 12.5°C i.e. with a wide range reflecting the soil temperature along the whole year. Excess air ranges from 0 to $0.02 cm^3 STP.g^{-1}$. The results show small variations with depth. Shallow piezometers tend to have low excess air and slightly higher recharge temperature while at 15 and 20m depth, excess air is slightly higher. However, a clear correlation with the hydrological regime could be observed. There is a global positive correlation between temperature and excess air, especially for samples from a recharge or discharge period. In general, samples with high NGT and low EA correspond at the end of recharge process (black squares; Fig 3; Figure2: March07, March08, and April09).

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3.4. Gas tracer distribution in groundwater

Table 1 summarizes the data obtained for CFC-11, CFC-12, SF_6 and $^3H/^3He$ per nest and per depth. All CFCs concentrations are relatively high in the aquifer, close to atmospheric maximum. CFC-11 concentrations range from 180pptv to 297pptv with an average concentration of 252pptv. These values indicate young groundwater (< 25 yrs of apparent age) in good agreement with the small size of this unconfined aquifer. The relative standard deviation (stdv/mean) is about 10% but tends to be lower for deep piezometer at the head of the catchment (piezometer F) indicating a higher stability of concentrations with depth. CFC-12 concentrations range from 241 to 571pptv with an average concentration of 492pptv. The relative standard deviation is slightly higher namely for H8 and H15 (Table 1). After correction from excess air, SF_6 concentrations are almost always above saturation with an average value of 7.4pptv (compared to atmospheric value of 6.99pptv in 2010). These high values most probably reflect SF_6 lithogenic production within the weathered part of the aquifer. This interpretation is supported by batch leaching experiments of the Kerrien alterites that showed high SF_6 fluxes derived from the rock (Aquilina et al., 2010).

Spatial distribution

CFC-11 and CFC-12 mixing ratios show similar vertical profiles for the different nests (Figure 4) suggesting that no clear evolution of the flow regime can be observed from upslope to downslope location of the catchment. CFC-11 vertical profile shows a decreasing concentration from 4 to 6m depth with the lowest concentrations observed at 6m depth. Then concentrations slightly increase up to 15m. These variations are the highest for nest H. At the opposite, nest J shows the most constant vertical profile both in CFC-11 and CFC-12. As the scope of the paper is to give a better knowledge of recharge processes, we will then focus on detailed variations at nest H. To understand these profiles variation, it is necessary to have a closer look on the data evolution with time.

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Temporal variation

The long term monitoring (Figure 2D) showed 2 clear trends in relation with the hydrologic condition. During the first dry cycles, in nest H, dissolved gases showed a general increase of their concentrations which is particularly pronounced from March06 to Sept06 (Figure 2D). Moreover, several data present value above the atmospheric maximum. The occurrence of these high value disappeared after 2007. During the humid period, from 2007 to 2010 concentrations stabilize more or less between present atmospheric concentration and maximum concentrations with a slight increase trend for the deepest piezometer (20 m). Two important decreases of CFC-11 values were observed during recharge periods of January 2007 and December 2008. In detail, this observation leads to distinguish vertical patterns of recharge and recession periods (Figure 5).

(1) During water table decrease (Apr and Jul06, March07, Apr09, Oct09 and March10), CFC-11 profiles were highly homogeneous with values ranging from 240 to 275pptv (Figure 5B). Variation of CFC-11 concentration at the same depth was found to be very low. Two campaigns (May05 and Sept06) showed values above atmospheric maximum. The origin of these high values is not fully understood yet.

(2) During recharge event, CFC-11 profiles showed lower values and a higher variability especially in the water table fluctuation zone (4 to 8m; Figure 5A). For instance, the particular recharge event of January 2007 is marked by a sharp decrease of CFC-11 concentrations for all nests and depth down to 15m, correlated to the fast water table increase and Cl^- decrease (Figure 2C). The lowest gases values were observed at 6m depth. The lack of gas data at 20m depth does not allow concluding on the maximum depth of influence. This concentrations fall was also noticeable for CFC-12 in a lesser extent, but not for CFC-113. A similar phenomenon was recorded during the recharge period of 2008 but to a lesser extent. This observation could explain the difference of global CFCs vertical profiles (Figure 4).

Apparent age and flow model distribution

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Mean apparent $^3H/^3He$ ages per depth range from 1.3 to 10.3 years with standard deviation ranging from 0.8 to 2.9 based on replicate samples (Table 1). Figure 6A compares apparent $^3H/^3He$ ages and CFC-11 ages derived from piston flow model. Although CFCs and $^3H/^3He$ ages present discrepancies, ages given by the 2 methods are in the same order of magnitude (<25yrs) and both present a general increase with depth. In detail, results show that for shallow depth down to 10m, most of CFC-11 provide similar or younger values than $^3H/^3He$, lower than 5yrs (Figure 6). On the contrary, for the 15-20m depth, CFC-11 apparent ages appear older than $^3H/^3He$ (about 20yrs for CFC-11 vs 5-12yrs for $^3H/^3He$) and a gap of CFC-11 ages can be observed between 12 and 18yrs. Part of this discrepancy can be due to the use of the piston flow model. At 15-20m, as apparent $^3H/^3He$ is about 10 years old, CFC data fall at the maximum of the atmospheric mixing curves, increasing uncertainty in apparent age determination. The systematic use of the rising part of the atmospheric curve could have biased old the apparent CFC piston ages.

Measured concentrations of CFC-11 and CFC-12 were compared with theoretical curves of different lumped parameter models (piston flow PFM, exponential mixing EMM, and binary mixing BMM; Figure 7). Data fit in the higher part of the curves again indicating young groundwater of less than 25yrs. Numerous points appeared out of the curves indicating discrepancies between the 2 tracers. The CFC-11 values above the PFM curve correspond to the 2 first years of sampling that showed data above the atmospheric maximum indicating CFC-11 contamination. In contrast, low CFCs concentrations obtained during the special recharge event of January 2007 (and December 2008 at a lower degree) are shifted down and plotted out of the model curves (Figure 7B). This is particularly visible for shallow depth (4,5-6m, Figure 7A). The other groundwater samples show mixed characteristics with data however mostly limited by a mixing line of present water and water recharge in 1988 (20yrs old). Taking into account the uncertainties, two main models explain most of the data: a piston model and an exponential mixing model.

We calculate CFC-11 MRT using an exponential model (EM). When no ages could be determined by EM (concentration higher than the maximum of the exponential curve), ages were derived from PFM. For

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young samples, similar ages were obtained with both models. For samples from 20m only exponential ages could be obtained. Finally, recession periods were better explained by piston flow model.

The resulted MRT is presented versus $^{3}H/^{3}He$ apparent age in Figure 6 B and C. A better agreement of CFC MRT and $^{3}H/^{3}He$ ages is obtained (Figure 6B), with most of CFC MRT below 12yrs, in better accordance with $^{3}H/^{3}He$ apparent age. In particular, MRT from recession periods were in good agreement with $^{3}H/^{3}He$ apparent age (Apr09, Oct09; Figure 6C). However, some CFC MRT, still appear older than $^{3}H/^{3}He$ apparent age, with no apparent correlation with depth. When looking at the hydrological period (Figure 6C), it appears that CFC MRT that are older than 3H/3He age correspond to the Dec08 recharge event. Similarly, CFC MRTs from recharge events of Jan07 appear still older than average apparent $^{3}H/^{3}He$ ages (about 15yrs even at shallow depth). For these events the type of lumped parameter model used to obtain MRT is not the origin of the discrepancy.

4. Discussion

4.1. Uncertainty in gas concentration estimation

The results show that ages obtained by $^{3}H/^{3}He$ and CFCs methods are in the same order of magnitude. Nevertheless, like in other studies combining these 2 age tracers, some discrepancies are observed (Happell et al., 2006; Kaown et al., 2009; Martin et al., 2016; Rademacher et al., 2001). Younger $^{3}H/^{3}He$ ages could be attributed to diffusive loss of ^{3}He to the atmosphere at times of low water table (Solomon et al. 1993, in Cook et al., 1995; Delbart et al., 2016; Martin et al., 2016; Rademacher et al., 2001). This would "reset" the $^{3}He/^{3}H$ clock inducing younger ages than CFCs ages (Cook et al., 1996). Noble gases are more prone to re-equilibration than CFCs because of their higher diffusivity (Han et al., 2006). This phenomenon could be enhanced by the thin vadose zone of the study site. In this study, CFCs retardation effect within the unsaturated zone would probably have low impact due to the low thickness of the

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permanently unsaturated zone (few meters thick) (Cook et al., 2006). In contrast, matrix diffusion processes between preferential flow path and matrix could cause discrepancy between the different tracers as they do not have the same aqueous diffusion coefficient (Cook et al., 2006). This phenomenon could explain why the apparent age derived from CFCs is higher than the water age derived from $^{3}H/^{3}He$. Finally, the discrepancy could result from equilibration temperature that induced uncertainty in CFCs age calculation with older CFCs ages if recharge temperature is set too high. In this study we found differences in recharged temperature derived from NGT and Mean Annual Air Temperature. Such discrepancy has already been reported by Hall et al. (2005). They show that the lower NGTs compared to MAAT could reflect potential degassing or the time necessary for the waters to equilibrate. This is probably the case for groundwater showing little excess air. In both cases as observed by Hall et al. (2005), NGTs may be influenced by hydrological conditions. Such influence may explain the correlation between excess air and recharge temperature as well as the sampling-time, i.e. hydrological conditions dependency. This result requires further research but supports the interpretation of gas data as tracers of recharge processes.

4.2. Influence of position along the hillslope

Contrary to studies in a similar context (Gooddy et al., 2006; Newman et al., 2010), gas concentrations do not indicate clear variations of flow regime down gradient through the catchment (Figure 4). This is consistent with the conclusion of a recent modeling study in a similar catchment (Molénat et al., 2013) where simulated CFCs groundwater mixing ratios were found not sensitive to the position along the hillslope in the range of the simulated MRT. In detail, the highest CFC vertical variations for nest H indicate a higher sensitivity to infiltration. Nest L appears to have similar behavior than nest H while nest J shows more homogeneous concentration with depth suggesting better permeability of the system. This slightly different physical property in nest J induces a lower influence of rapid infiltration which supports previous observations on spatial variation of water quality along the hillslope (Rouxel et al., 2011).

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4.3. Influence of hydrologic condition on CFCs variability

At short time scale, the results show the influence of the sampling period on CFCs variability in groundwater. During low flow, there are small variations between different sampling periods (example of March07, March10 and Apr09) proving that one single time field sampling can be sufficient in most of cases. However, in this shallow aquifer, results show that recharge events have a strong impact on CFCs concentrations.

Processes influencing concentrations during beginning and end of recharge

The most representative example is the sampling during the recharge event of January 2007 and the following discharge period of March 2007. This sampling was done during a high recharge event after a dry period. At the beginning of the recharge period, diluted Cl^- concentrations reflect the infiltration of highly diluted rainwater through preferential flow path in the vadose zone. The dilution is observed on Cl^- down to 10m. As a consequence, vertical Cl^- concentrations in groundwater reflect a mixing between permanent groundwater (35mg/l) and rainwater (about 10mg/l; Figure 8B).

In Kerrien catchment, as the "old" permanent groundwater have similar CFC-11 mixing ratio than the atmosphere at the date of sampling (249pptv; Figure 8A), mixing of permanent groundwater with rainwater should lead to a vertical homogeneous CFC-11 concentration close to the atmospheric concentration (Figure 8A). Instead, CFC-11 concentrations show lower value with a sharp decrease between 6 and 8m depth. The depleted values most probably reflect CFC-11 degradation as described in several studies for groundwater with anaerobic conditions (Cook et al., 2006, 1995; Darling et al., 2012; IAEA, 2006; Oster et al., 1996). This hypothesis is reinforced by the low DO values measured for sampling in January 2007 (4.5mg/l). The maximum degradation is observed at 6-8 m depth which corresponds to the water table level during low flow. This observation is in good agreement with previous study that mentioned a denitrification zone down to 10 m in Kerrien catchment (Legout et al., 2005; Martin et al., 2006).

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Before recharge, during low flow period, groundwater level is below 7m depth. In the fluctuation zone, it is assumed that residual matrix water undergoes a chloride enrichment and CFC-11 degradation. During recharge, as water level rise in the fluctuation zone, permanent groundwater mixes with residual matrix water enriched in Cl^- and SF6 and depleted in CFC-11. In the same time, major infiltration of rainwater induces a decrease of Cl^- and limits the influence of CFC-11 degradation at shallow depth.

The major dilution effect on Cl^- down to 6m reflects preferential flow paths and thus the heterogeneity of the aquifer. The contribution of the matrix water is supported by the increase of SF_6 values (up to 11.4pptv at 10m vs 6.3pptv in the atmosphere in 2007). These SF_6 values in groundwater above atmospheric maximum reveal in situ production and confirm contribution of pore water with a higher contact time. Similar processes occur for the recharge event of December 2008 but to a less extent probably due to a lower water level rise during December 2008. Moreover the dry conditions during the 2 years before January 2007 are supposed to have enhanced the CFC-11 degradation process leading to undatable water samples.

At the end of recharge, (March07 or Apr09), vertical rainwater contribution has stopped as well as CFC-11 degradation since the system has been re-oxygenated. Shallow piezometers show high Cl^- concentrations. A decreasing influence of matrix contribution with depth explains the vertical Cl^- profile in nest H. During this recession period, except at 5-6m depth, CFC ages can be estimated using a piston flow model (Figure 6). CFC-11 profile shows a downward evolution with a modern value at the top of permanent saturated zone and an older value at the bottom in good agreement with the $^3H/^3He$ age profile. Similar results were obtained for April 2009. Such evolution indicates that mixing with matrix water occurs mainly at shallow depth down to 6-8m, and then horizontal flows dominate in the system.

This interpretation is strengthened by the NGT and EA computation (Figure 3). At the end of the recharge process, when Cl^- concentration is the highest, groundwater is characterized by relatively high NGT and lower EA. Such evolution in water chemistry can be explained by the contribution of pore water from the soil and the vadose zone. Although the potential mechanism remains unclear, such pore water closely bounded to the rock matrix would exclude air inclusion. During the beginning of the recharge process,

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most of the NGT show low value below 12°C associated to relatively high EA which could reflect the rapid increase of water table due to the by-pass of "cold" rainwater.

Estimation of contribution from the different end-member:

To validate the processes previously described, we did a first calculation using a 3 end-members mixing model: permanent groundwater, pore water and rainwater (Table 2). For rainwater, we used the atmospheric value of CFC-11 in 2007 and measured Cl^- value. For the permanent groundwater, we use CFC-11 value of F and Cl^- values in H15 during March 2007. The highest measured Cl^- concentration (49.4mg/l in H6) was assumed to represent the Cl^- concentration in matrix water. Then we assumed a total degradation of CFC-11 leading to value close to 0pptv.

The observed Cl^- and CFC-11 values at 6m were well simulated. During recharge period, a similar contribution of permanent groundwater and rainwater is found of about 40%. Matrix contribution is estimated to be up to 15%. At the end of recharge, rainwater contribution falls to 0. Shallow piezometers show high Cl^- concentrations that can be reproduced assuming a contribution of 50% of matrix water to permanent groundwater. Although, this computation can be only considered as a rough estimate, it confirms the influence of pore water contribution and indicates that vertical transfer are not negligible in the upper 6-8 of the groundwater section. A specific modeling should be carried out to further investigate the amplitude of this mechanism and potential effect in larger aquifers.

This interpretation shows that *i*) CFCs do not show a constant evolution during the hydrologic cycle and *ii*) different conceptual flow models have to be considered depending on depth and sampling period. During recharge, vertical flow and mixing dominate, limiting the use of apparent ages. In contrast, during recession periods, horizontal flows dominate and ages can be derived from piston classical flow models. At greater depth (20m), tracer ages are obtained using exponential model showing that horizontal flows dominate and recharge event has low influence. This is consistent with previous results in similar context where the decrease in apparent ages below 10m was attributed to groundwater moving horizontally through a second flow system (Cook et al., 1996). As a consequence, in such systems, the use of these

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classical models to estimate groundwater ages from atmospheric tracers should be limited to recession periods. This long-term monitoring shows that a good confidence can be given to apparent age at this period, with a difference of 2.8 yrs (+/- 2.9) for the 8-10 m depth interval for 3 samplings with similar hydrological conditions (March07, Apr09 and March10). It is worth noting that the impact of the recharge period tends to disappear rapidly.

4.3. Temporal distribution of Mean Residence Time

In the previous sections, we mainly focused on one hydrological cycle to understand groundwater mechanisms. Here we examine the tracer ages obtained at each depth for all the investigated period (Figure 9). Recharge samples (Jan07 and Dec08) presenting CFCs degradation were not used here. A general increase of MRT with depth is observed. It corresponds to an overall vertical groundwater velocity of 0.5 to 0.7m per year, which is in good agreement with values obtained from previous tracing experiments in this system (Legout et al., 2007) and studies in similar geological context (Cook et al., 1996). However, a vertical layering of the groundwater MRTs is observed (Figure 9). In the 0 to 6m depth piezometers, i.e. within the water table fluctuation zone, very short to a few years MRT are observed. In the 8 to 10m depth piezometers, the MRT range from a few years to a decade, while they increase to 10 to 15 years in the 15 to 20m depth piezometers. This layering corresponds to the Cl^- vertical profile which shows a decreasing variability of concentration with depth (Rouxel et al., 2011).

The long term monitoring showed the effect of annual recharge on the residence time of the groundwater. Present water only appears just after recharge periods very close to the water table. During the recession periods, even at the water table surface, the MRT is already close to 2.5 years (March05, Apr09) and water shows high Cl^- concentrations which confirms the influence of long-term residence time pore water from the unsaturated zone. At 8-10m during the dry period (2005-2006 cycle), we show that groundwater is rejuvenated as water level increases. During recession, residence time tends to increase up to about 12 years as less rainwater enters in the system. The January 2007 special event highly modifies the system since a high amount of young water enters the system. In consequence, during the humid period, MRT are

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much younger with a mean value about 5yrs. At 15m depth the MRTs are more stable than above. However the influence of the strong recharge event of January 2007 can still be observed: residence time is about 10-12yrs before January but decreases after January and stabilizes at 8yrs during 2008 and 2009 cycles.

At 20m, during the dry period, a slight increase of residence time is observed in relation with a slight Cl^- increase. Then, residence times decrease from June 2007 to April 2009 and stabilize at 8yrs indicating that young water rejuvenates slowly the system at this depth. This rejuvenation explains the observed decrease of Cl^- concentration since January 2007 (Figure 2). This observation confirm that groundwater at 20m behaves relatively independently from other depths.

4.4. Consequences for conceptual model of the catchment

The highly similar pattern observed for both CFCs and Cl^- concentrations and the synchronous variations during recharge at depth of 5 to 15m highlight the occurrence of vertical transfer from the variably saturated zone (2 -6m) down to 10-15m depth. These results contradict the usual consideration of a major horizontal transfer from up to downslope as assumed by the approximation of Dupuit-Forsheimer which neglects vertical flows. These results also confirm that a double flow system can explain temporal variation in anionic concentrations as suggested previously (Legout et al., 2007; Rouxel et al., 2011). In addition, this work gives a better understanding of the mechanism controlling recharges. Local vertical downward flows occur during high recharge period inducing temporal variability in the anionic concentration of the permanent groundwater. High CFC-11 degradation at shallow depth (and CFC-12 to a lesser extent) indicates that during infiltration, rainwater mixes with evolved water from the matrix. As contribution of rainwater decreases, evolved matrix water contribution increases leading to Cl^- increase. This phenomenon is observed only when a sufficient groundwater level is reached and is enhanced by previous dry conditions.

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5. Conclusions

Using a detailed analysis of long-term time-series of dissolved gas tracers in a shallow aquifer, we investigated how tracer concentrations and associated residence times evolve with depth and which flow processes (by-pass/ PFM/EM/Mixing) explain these evolutions. The long-term monitoring of dissolved gases shows a good repeatability in the permanent saturated zone and more variable apparent ages in the water fluctuation zone dominated by mixing. CFCs and $^{3}H/^{3}He$ indicate young groundwater. SF_6 show concentrations above the mean atmospheric mixing ratio indicating lithological production which limits its utilization in this aquifer for groundwater dating.

In this study, we show that hydrologic conditions have more influence on dissolved gas concentration than the location along the hillslope. Two different flow processes have to be considered between recharge and recession periods. During water table decreases, CFC-11 vertical profiles are highly homogeneous. During recharge periods, CFC profiles show variable but lower CFCs values reflecting contribution of modern rainwater and of the weathered matrix with variable proportion at the beginning of recharge and at the end of recharge period. This conceptual model is consistent with Cl^- concentration as well as excess air and recharge temperature evolution during the hydrologic cycle. The small variations observed between the sampling periods during recessions show that a single field sampling during groundwater recession period can be sufficient in most cases to estimate MRT. A general increase of mean residence time with depth is observed at longer time scales. In addition, this study shows a clear stratification between the first meters below the water table which exhibit highly variable gas concentrations, and the deep part of the aquifer which is much more homogeneous. Such stratification indicates that hydrological processes strongly differ in the variably saturated zone which acts as a buffer. These results confirm the interest of CFCs as age dating tracers for shallow groundwater, but also as natural tracers to conceptualizing flow processes. In this way, this work confirms the interest of atmospheric tracers to better constrain conceptual hydrological models regarding to their ability to represent different processes according to the hydrological period. Finally, this work demonstrates the

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interest of long-term monitoring of dissolved gases in complex hydrogeological environment for a better understanding of the recharge processes, with specific recharge events that impact groundwater over several cycles.

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Table caption

Table 1. Number of sample (No), mean values and standard deviation (STDV) obtained for CFC-11, CFC-12, SF_6 and $^3H/^3He$, per nest and per depth. For CFCs and SF6, both mixing ratio and apparent tracer ages are given. The relative standard deviation (RSD) is also given for CFCs and SF_6.

Table 2. End-members values and their respective contribution in H6 for a recharge period (January 2007) and an end of recharge period (March 2007).

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Figure captions

Figure 1. (A) Topography of the Kerrien catchment with (B) a schematic view of the transect F, H, J, L and the piezometer nests.

Figure 2. Temporal evolution of precipitation, hydrogeology and hydrochemical parameters (Cl^-, CFC-11, SF6) of the system.

(A) Annual precipitation deviation in comparison to the mean interannual precipitation over the year 2005 to 2010; (B) groundwater depth time series from the soil surface at nest H (upslope), J (midslope) and L (downslope), precipitation and potential evapotranspiration at the soil surface time series. Sampling periods have been reported on the water level and water level variation in H (white circle). (C) Cl- concentration ($mg.l^{-1}$) and (D) CFC-11 and SF6 dissolved concentrations (pptv) respectively measured from January 2005 to October 2010 at different depth in nest H (upslope) and F compared to daily water level variation (in cm.d-1) in H15. (E) SF6 dissolved concentrations (pptv) respectively measured from January 2005 to October 2010 at different depth in nest H (upslope) and F. The dash lines (D-E) show the maximum mixing ratio in the atmosphere (black line) and the present mixing ratio (grey line). Main recharge period have been highlighted in grey.

Figure 3. Recharge temperature and excess air derived from gas concentrations

Figure 4. Vertical profiles of CFC-11 and CFC-12 concentration at nests H (upslope), J (midslope) and L (downslope) (A) and corresponding apparent age evolution (B). Apparent $^3H/^3He$ age vertical profile has also been reported (B). Bars indicate the coefficient of variation between the different sampling.

Figure 5. Comparison of CFC-11 vertical profiles (y-axis unit is m) for (A) recharge period; and (B) recession period. Sampling periods with less than 3 depths have not been represented. Dashed line represent equilibrium concentration in 2010; maximum equilibrium concentration is 274pptv (1994).

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Figure 6. Comparison of $^{3}H/^{3}He$ and CFC-11 apparent ages at different depths, assuming a piston fow model for CFCs ages (A). Considering the discrepancy at 15-20m, CFC-11 ages have been determined using the best fitted model between EM or PFM and compared to 3H/3He apparent ages according to depth (B) and the period of sampling (C).

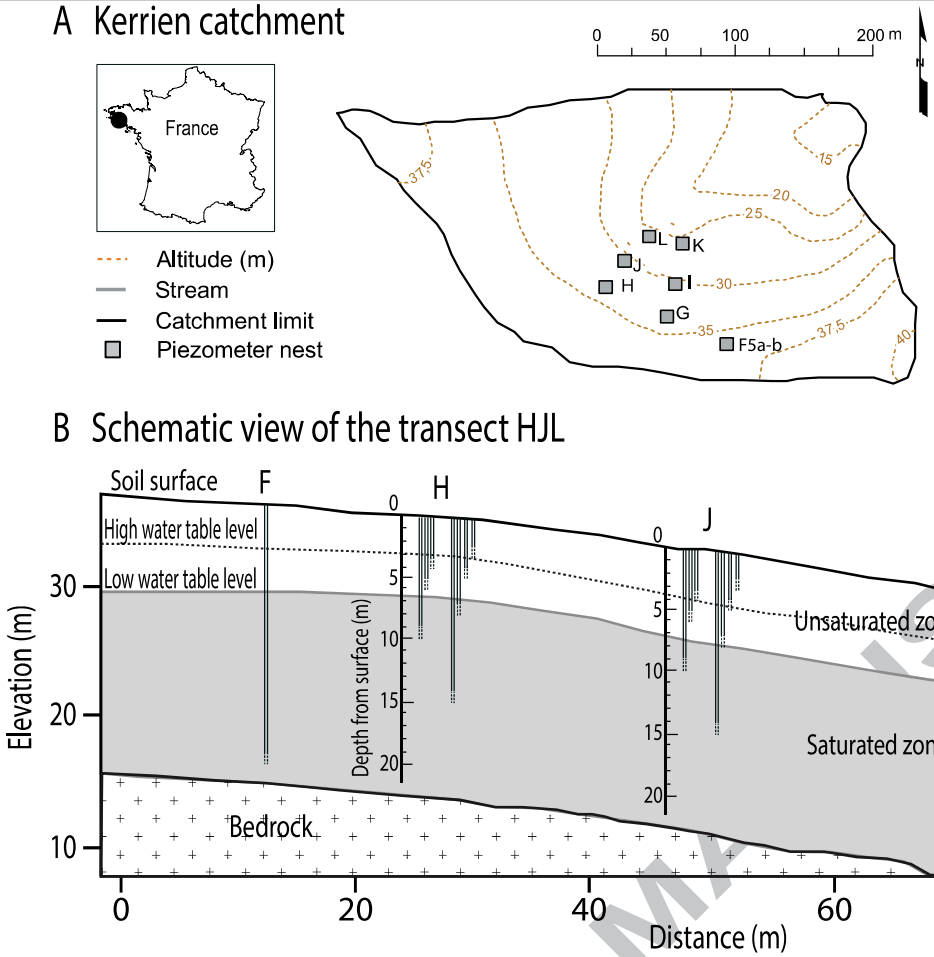
Figure 7. CFC-11 and CFC-12 dissolved concentrations compared to flow models (Piston Flow Model (PFM); Exponential Model (EM) for sampling in 2005, 2007 and 2010; Binary Mixing Model (BMM)) depending on depth of sampling for all transect (A) or only plot H (B) or depending on the hydrological period in plot H: Recharge (C); Recession (D). During recharge, data are scattered below the model line.

Figure 8. CFC-11 (A) and Cl^- (B) profile in nest H during recharge event of January07 and following recession period (March07). Rainwater and permanent groundwater end-member are indicated by rectangles and their mixing by the arrow. Solid black line with crosses show the atmospheric mixing ratio of CFC-11 during the past 20 years.

Figure 9. Time series of ages derived from CFC-11 and 3H/3He for the unsaturated zone (4-6m) and the permanent groundwater (8m, 10m, 15m, 20m).

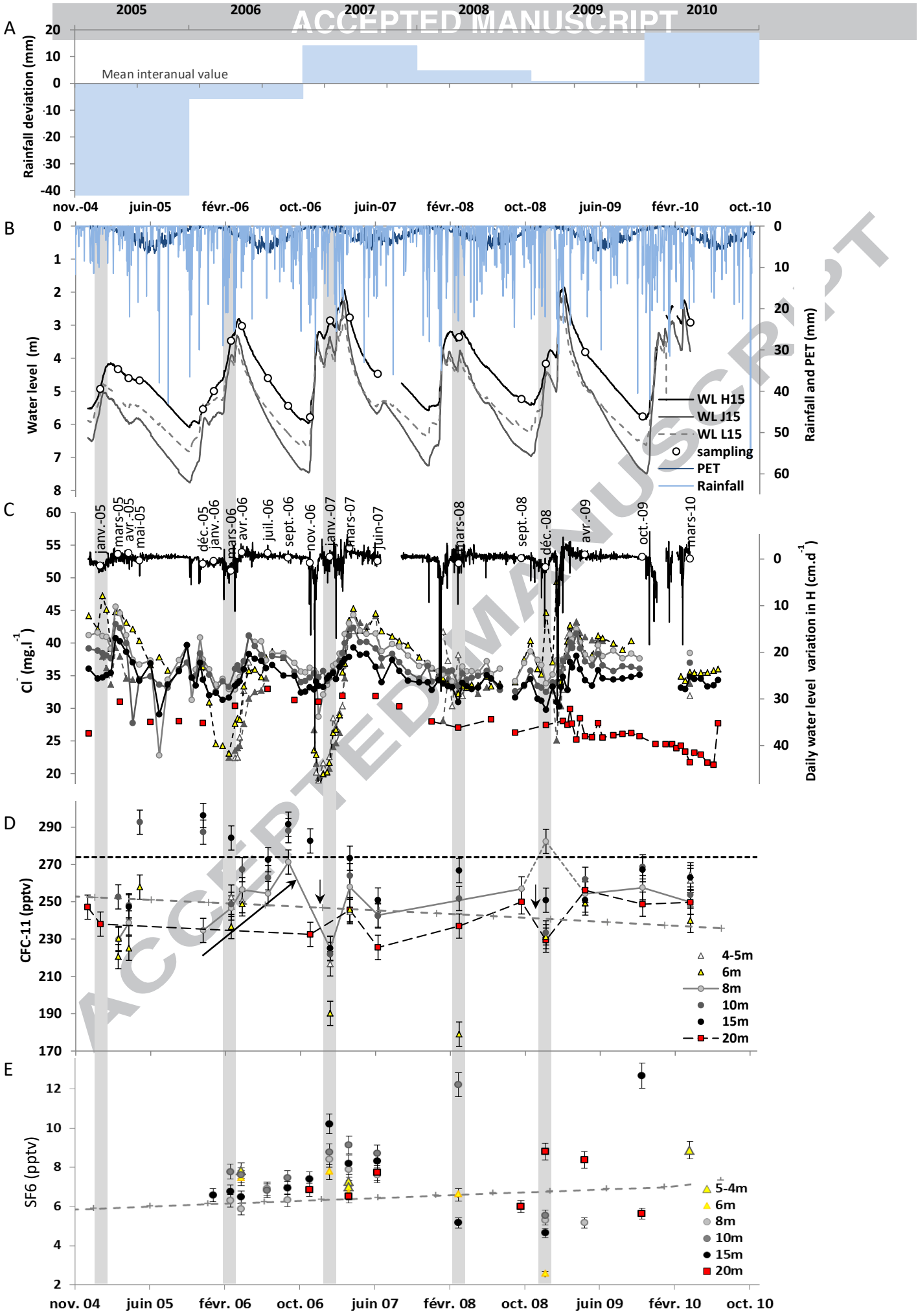
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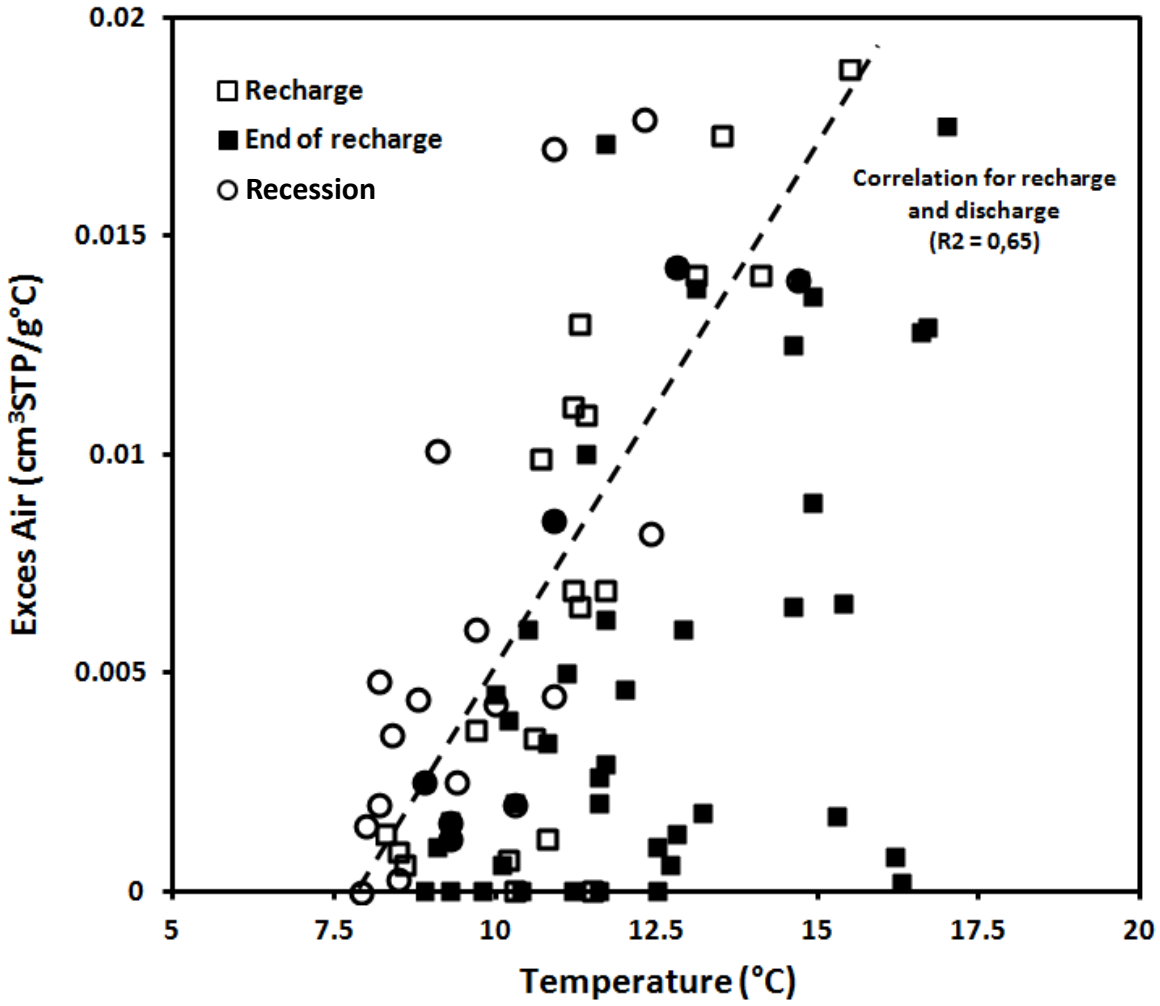
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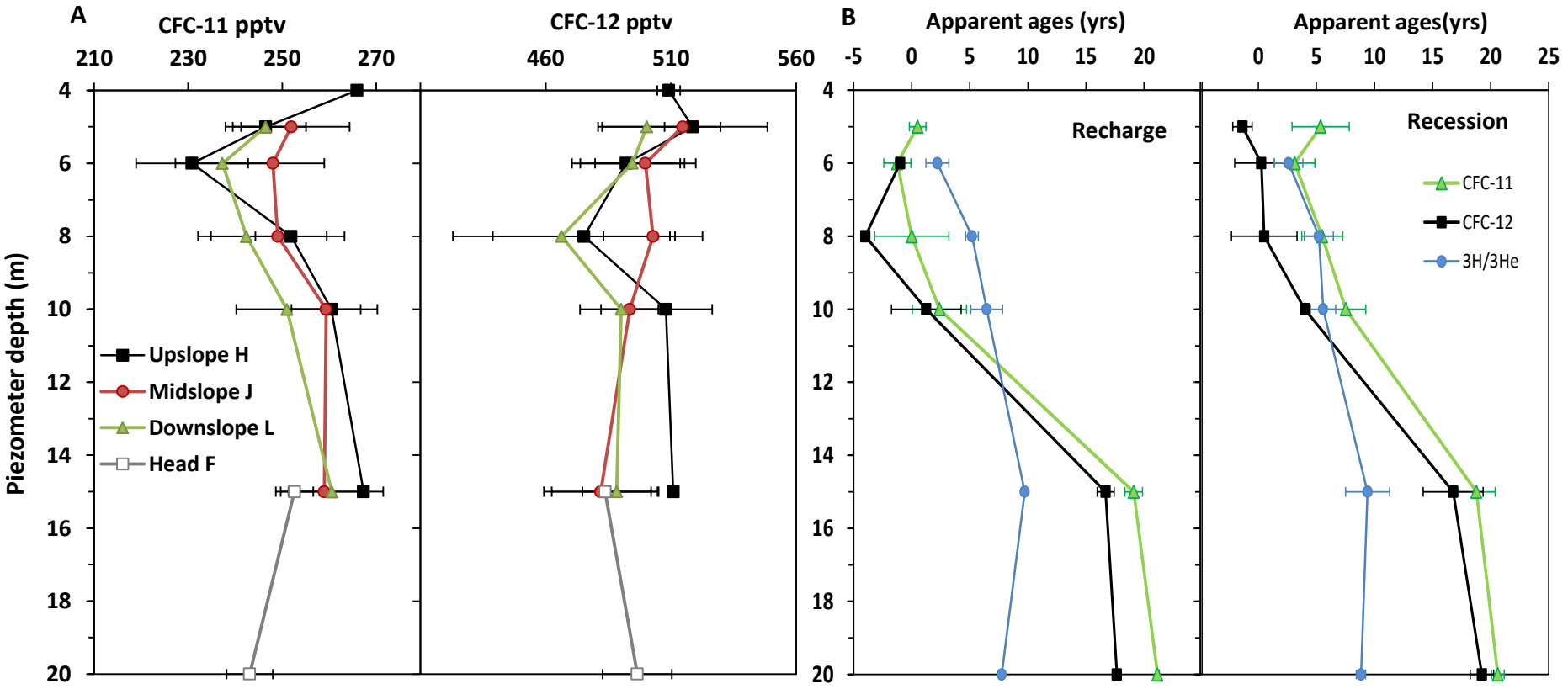


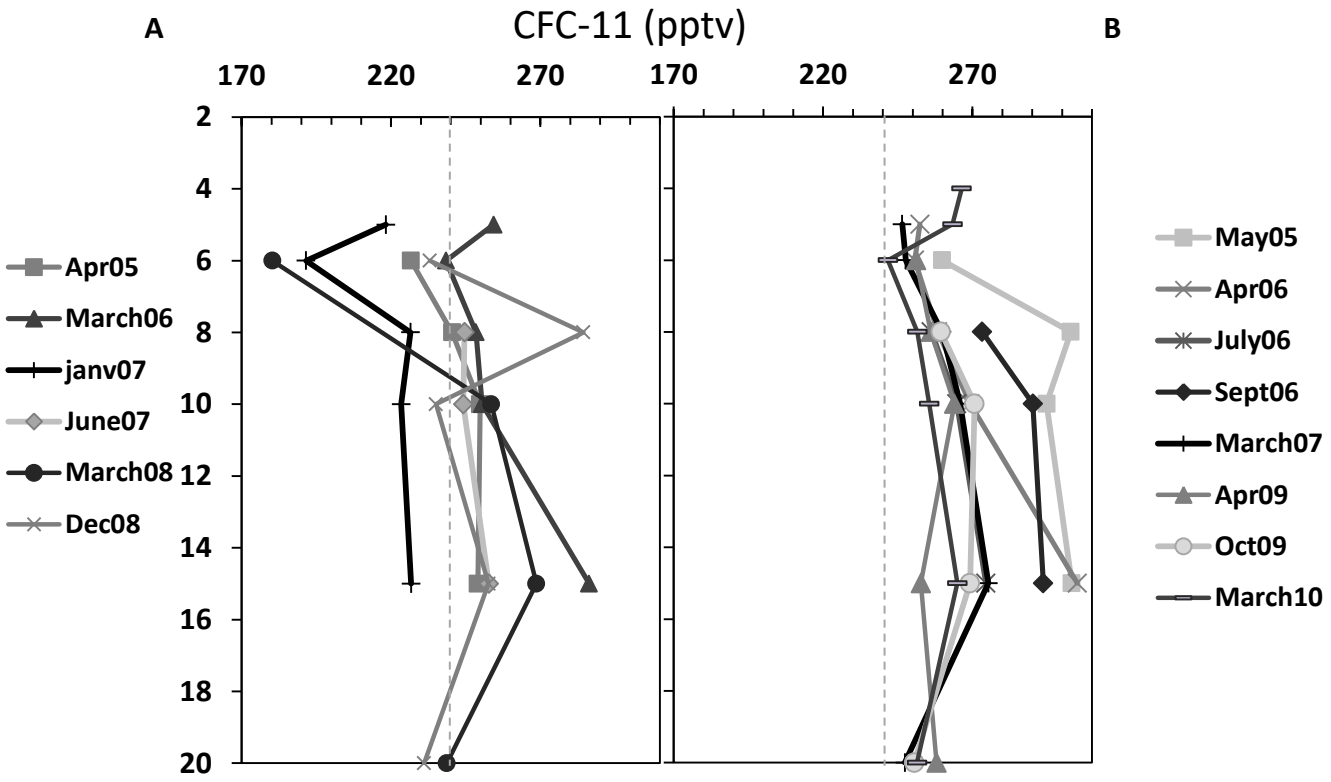
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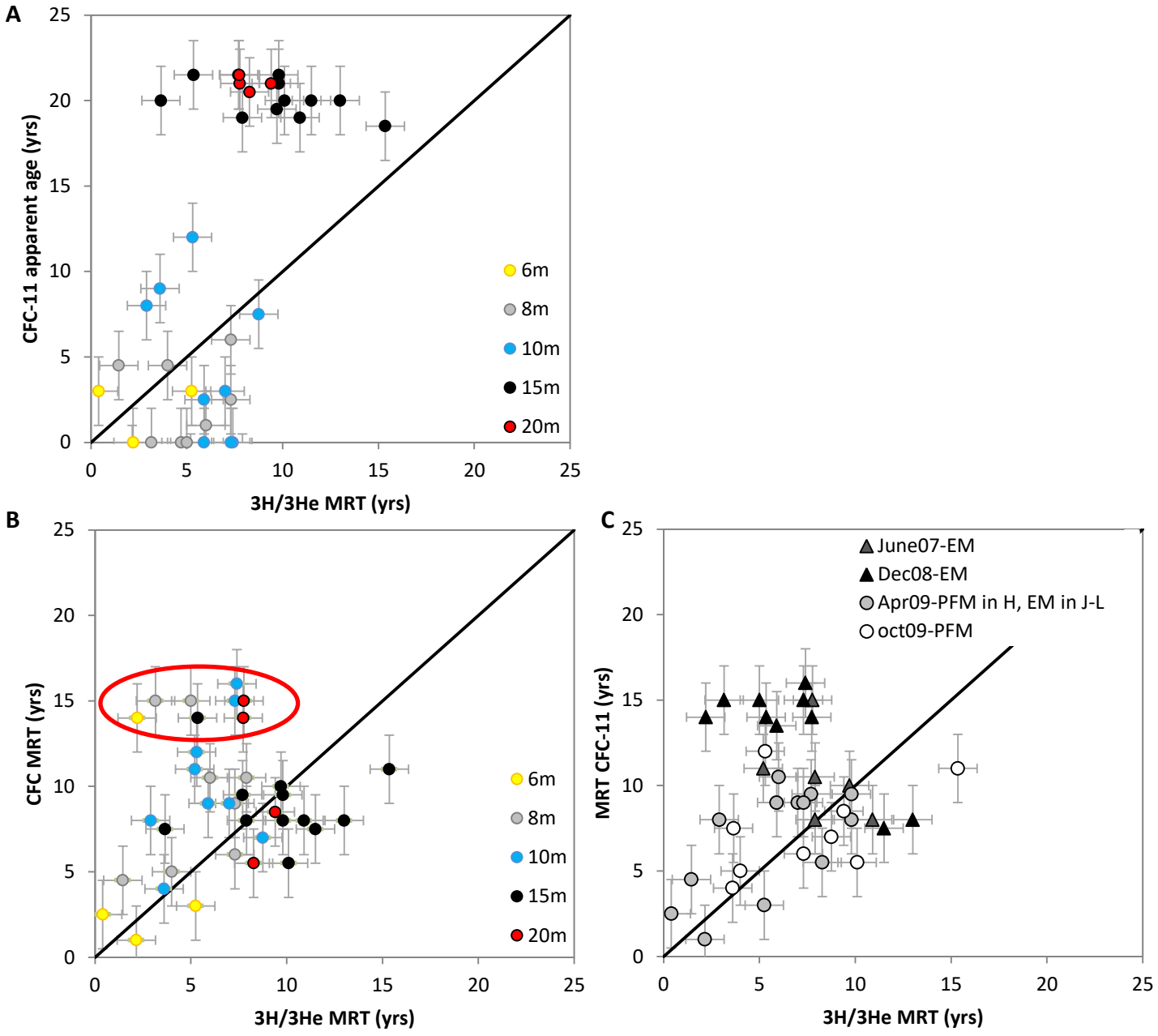
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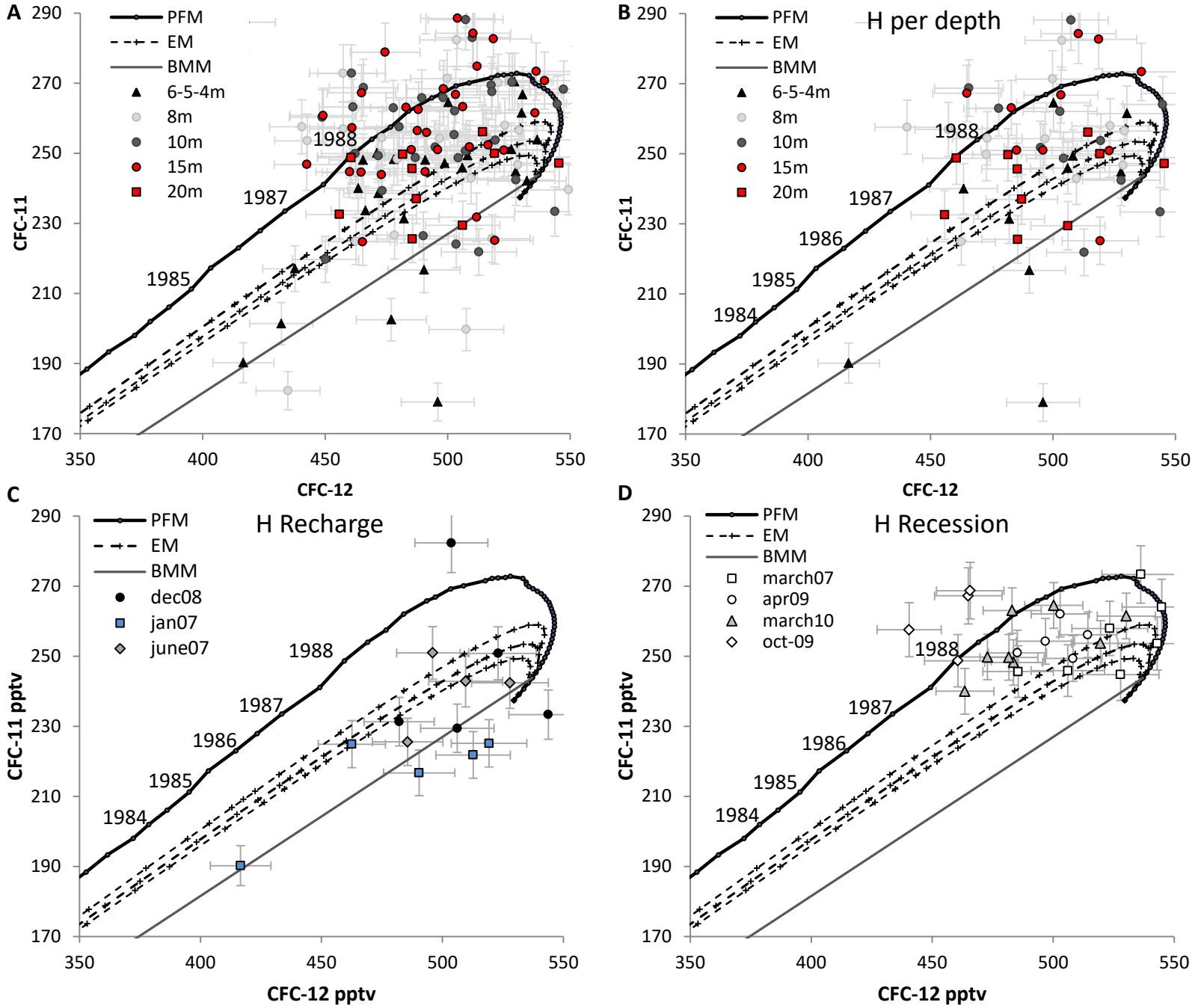


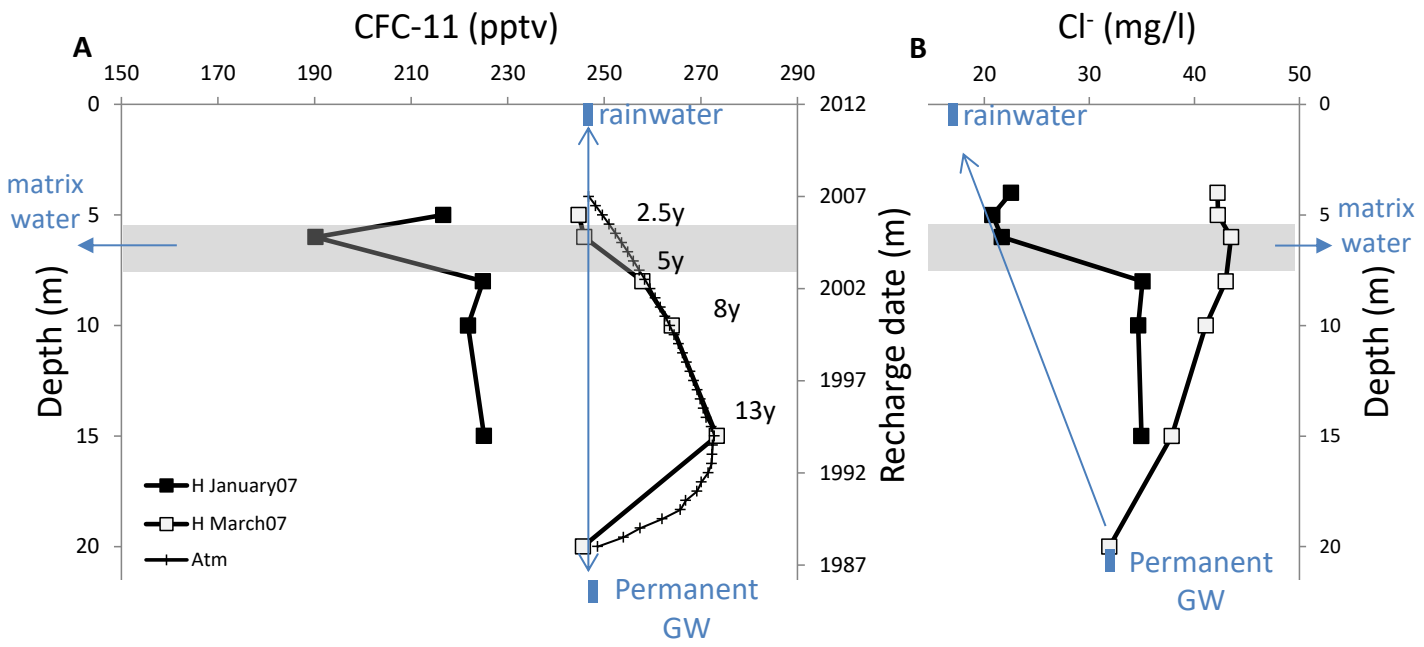

0.02
0.015
0.01
0.005
0
Exces Air (cm³STP/g°C)
5
7.5
10
12.5
15
17.5
20
Temperature (°C)
Recharge
End of recharge
Recession
Correlation for recharge and discharge (R2 = 0,65)

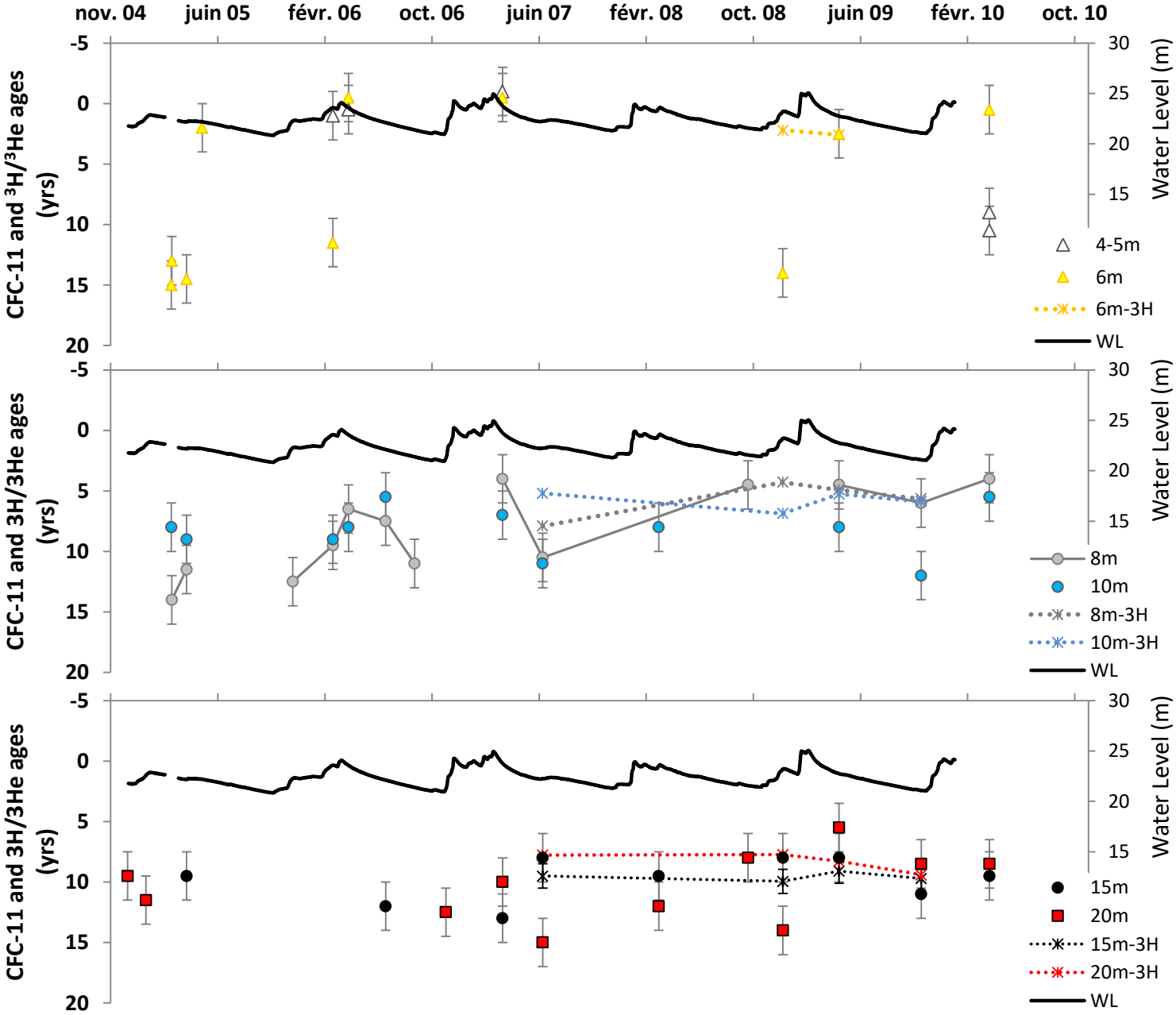

A
CFC-11 pptv
210
230
250
270
CFC-12 pptv
460
510
560
B
Apparent ages (yrs)
-5
0
5
10
15
20
Apparent ages(yrs)
0
5
10
15
20
25
Piezometer depth (m)
4
6
8
10
12
14
16
18
20
Upslope H
Midslope J
Downslope L
Head F
Recharge
Recession
CFC-11
CFC-12
3H/3He


A
CFC-11 (pptv)
B
170
220
270
170
220
270
2
4
6
8
10
12
14
16
18
20
Apr05
March06
janv07
June07
March08
Dec08
May05
Apr06
July06
Sept06
March07
Apr09
Oct09
March10


A
CFC-11 apparent age (yrs)
3H/3He MRT (yrs)
6m
8m
10m
15m
20m
B
CFC MRT (yrs)
3H/3He MRT (yrs)
6m
8m
10m
15m
20m
C
MRT CFC-11 (yrs)
3H/3He MRT (yrs)
June07-EM
Dec08-EM
Apr09-PFM in H, EM in J-L
oct09-PFM


A
PFM
EM
BMM
6-5-4m
8m
10m
15m
20m
1988
1987
1985
CFC-11
CFC-12
290
270
250
230
210
190
170
350
400
450
500
550
B
H per depth
1988
1987
1986
1985
1984
C
H Recharge
dec08
jan07
june07
CFC-11 pptv
CFC-12 pptv
D
H Recession
march07
apr09
march10
oct-09


A
CFC-11 (pptv)
150
170
190
210
230
250
270
290
0
5
10
15
20
Depth (m)
2012
2007
2002
1997
1992
1987
Recharge date (m)
rainwater
matrix water
2.5y
5y
8y
13y
H January07
H March07
Atm
Permanent GW
B
Cl- (mg/l)
20
30
40
50
rainwater
matrix water
Permanent GW
Depth (m)


nov. 04
juin 05
févr. 06
oct. 06
juin 07
févr. 08
oct. 08
juin 09
févr. 10
oct. 10
CFC-11 and ³H/³He ages (yrs)
Water Level (m)
-5
0
5
10
15
20
30
25
20
15
4-5m
6m
6m-3H
WL
CFC-11 and 3H/3He ages (yrs)
Water Level (m)
8m
10m
8m-3H
10m-3H
WL
CFC-11 and 3H/3He ages (yrs)
Water Level (m)
15m
20m
15m-3H
20m-3H
WL

		CFC-11						CFC-12						SF_6				$^3H/^3He$		
Atmospheric maximum		272.8						546.32						6.99						
		mixing ratio (pptv)				App. Age (Years)		mixing ratio (pptv)				App. Age (Years)		mixing ratio (pptv)				App. Age (Years)		
Depth	Name	**No**	Mean	STDV	RSD	Mean	STDV	**No**	Mean	STDV	RSD	Mean	STDV	**No**	Mean	STDV	RSD	**No**	Mean	STDV
	F	**19**	247.1	10.1		19.3	2.1	**16**	491.7	33.2		17.6	3.6	**11**	6.4	1.5		**4**		
15	F5a	**8**	252.6	7.9	3%	18.4	2.4	**6**	483.7	42.7	9%	16.8	5.4	**4**	5.1	0.9	18%			
20	F5b	**11**	243.1	9.9	4%	20.0	1.6	**10**	496.5	27.5	6%	18.2	1.8	**7**	7.1	1.2	17%	**4**	8.3	0.8
	H	**63**	253.3	22.5		6.8	7.5	**51**	498.9	47.3		9.4	8.6	**38**	7.4	1.9		**14**		
5	H5	**5**	246.5	17.1	7%	2.8	4.6	**3**	518.6	22.4	4%	-1.3	2.5	**2**	7.6	0.4	5%			
6	H6	**12**	230.8	23.9	10%	0.4	2.5	**7**	492.1	43.2	9%	3.5		**5**	6.2	2.1	34%	**2**	1.3	1.3
8	H8	**15**	251.9	15.2	6%	3.2	4.2	**13**	475.3	72.8	15%	-3.8	0.4	**9**	6.7	1.2	17%	**4**	5.3	2.9
10	H10	**16**	260.5	19.5	8%	4.6	4.9	**13**	508.0	37.0	7%	2.6	3.8	**9**	8.2	1.9	23%	**4**	4.8	1.3
15	H15	**14**	267.3	19.7	7%	18.0	3.2	**13**	510.9	25.4		16.0	3.1	**11**	7.6	2.3		**4**	12.3	2.4
	J	**51**	255.1	19.8		9.3	7.8	**46**	493.3	39.4		14.6	7.6	**39**	7.4	1.7		**10**		
5	J5	**4**	251.9	24.9	10%	7.0	6.6	**3**	514.7	67.6	13%	-2.5		**3**	7.9	0.9	12%			
6	J6	**7**	248.0	21.9	9%	4.3	4.8	**6**	499.8	40.2	8%	-3.0		**5**	7.0	0.7	11%	**1**	5.3	
8	J8	**10**	249.1	28.5	11%	5.4	7.0	**8**	502.8	39.5	8%	4.5		**7**	7.4	1.9	26%	**2**	6.2	1.6
10	J10	**15**	259.3	14.7	6%	5.3	4.8	**13**	493.5	22.9	5%			**11**	7.6	2.4	32%	**3**	6.0	2.1
15	J15	**15**	258.9	15.3	6%	19.1	2.0	**16**	481.9	45.3	9%	17.4	3.7	**13**	7.4	1.4	18%	**4**	6.2	2.0
	L	**41**	249.9	21.4		9.1	8.1	**35**	485.7	49.0		14.5	7.9	**24**	7.7	2.2		**11**		
5	J5	**3**	246.5	10.5	4%	3.3	0.4	**3**	500.4	35.8	7%	-0.5		**1**	8.2					
6	J6	**5**	237.3	20.1	8%	1.1	2.0	**5**	494.6	41.5	8%	-1.0		**3**	5.8	1.7	30%	**1**	2.2	
8	J8	**9**	242.4	20.6	8%	4.9	2.5	**8**	466.3	86.7	19%			**4**	7.7	3.5	46%	**3**	4.4	1.5
10	J10	**12**	251.0	21.5	9%	5.1	5.2	**9**	490.2	33.0	7%			**8**	7.4	2.4	33%	**3**	7.3	1.4
15	J15	**12**	260.6	21.8	8%	19.3	2.1	**10**	488.3	27.4	6%	17.6	3.7	**8**	8.7	1.3	15%	**4**	10.3	0.8
	Total	**174**	252	20.5		9.9	8.2	**148**	493	43.9		13.5	7.8	**112**	7.4	1.9		**39**	6.8	3.2

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	Permanent groundwater	Matrix Water	Rainwater	H6 recharge Jan07	H6 end of recharge March07
Cl^- (mg/l)	35	50	8	25.1	42.5
CFC-11 (pptv)	249	0	247	210.75	261.5
Contribution (Recharge)	40%	15%	45%		
Contribution (End of recharge)	50%	50%	0%		

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- We analyzed long-term time-series of dissolved gas tracers in groundwater
- We investigated recharge processes through time and depth using CFCs and $^{3}H/^{3}He$ according to hydrological conditions
- Different flow processes occur during recharge and recession periods
- Hydrological processes differ between the first meters and the deeper part of the aquifer

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