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Article

Co-binding of pharmaceutical compounds at mineral surfaces: Molecular investigations of dimer formation at goethite/water interfaces

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1	Co-binding of pharmaceutical compounds at mineral
2	surfaces: Molecular investigations of dimer formation at
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Abstract

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The emergence of antibiotic and anti-inflammatory agents in aquatic and terrestrial systems is becoming a serious threat to human and animal health worldwide. Because pharmaceutical compounds rarely exist individually in nature, interactions between various compounds can have unforeseen effects on their binding to mineral surfaces. This work demonstrates this important possibility for the case of two typical antibiotic and anti-inflammatory agents (nalidixic acid (NA) and niflumic acid (NFA)) bound at goethite (α -FeOOH) used as a model mineral surface. Our multidisciplinary study, which makes use of batch sorption experiments, vibration spectroscopy and periodic density functional theory calculations, reveals enhanced binding of the otherwise weakly bound NFA caused by unforeseen intermolecular interactions with mineral-bound NA. This enhancement is ascribed to the formation of a NFA-NA dimer whose energetically favoured formation (-0.5 eV compared to free molecules) is predominantly driven by van der Waals interactions. A parallel set of efforts also showed that no co-binding occurred with sulfamethoxazole (SMX) because of the lack of molecular interactions with co-existing contaminants. As such, this article raises the importance of recognising drug co-binding, and lack of co-binding, for predicting and developing policies on the fate of complex mixtures of antibiotics and anti-inflammatory agents in nature.

Introduction

Thousands of different emerging pharmaceutical contaminants occur in soils, groundwater, surface waters as well as seawater from human and intensive farming activities. Antibiotics and anti-inflammatory agents in terrestrial and aquatic environments, in some instances at levels as high as several hundred ng per L ³⁻⁶ are posing detrimental ecological and health effects especially because of their growing use in human and veterinary medicine. Because the fate of these compounds is often tied to their affinities to surfaces of soil and sediment mineral particles ^{7,8}, adsorption through synergistic drug interactions is likely to become an emerging mechanism in contaminated environments.

Although contaminants rarely exist in isolation, they often have been studied individually with respect to sorption and/or complexation with naturally occurring minerals. 9-12 Sorption of individual compounds to environmental surfaces involves different mechanisms including metal bond, hydrogen bond, and van der Waals interactions ¹³. In multicomponent systems, co-existing contaminants can compete for surface binding sites, or cooperatively bind by co-neutralisation of surface charge and/or by direct molecular interactions. While competitive adsorption has been widely investigated ^{14–16}, cooperative effects have been never reported for widely used antibiotic and anti-inflammatory agents. In addition, because most traditional environmental models are based on an individual contaminant basis, little is known on their fate in mixed contaminant systems.

In this work, we assessed the ability of three typical antibiotic and
anti-inflammatory agents detected in affected environments ^{6,17} (nalidixic acid (NA),
niflumic acid (NFA) and sulfamethoxazole (SMX)) to co-bind at minerals surfaces.
Goethite (α -FeOOH) is selected as model mineral because it is one of the most stable
thermodynamically iron oxyhydroxide at ambient temperature and the most abundant
one in natural settings. NA is a quinolone antibiotic that is widely used in humans and
animals and that typically co-occurs with SMX, a sulfonamide antibiotic commonly
used to treat a variety of bacterial infections ¹⁸ . Niflumic acid (NFA) is a non-steroidal
anti-inflammatory that is often used for rheumatoid arthritis. 19 As will be detailed in
this work, investigations are mainly focused on NA and NFA as no co-binding effect are
typically observed with SMX. Vibration spectroscopic and density functional theory
(DFT) calculations of NA/NFA co-binding, as well as batch kinetic, pH-edges and
isotherms, were used to resolve uptake mechanisms of NA and NFA at goethite in
isolated vs. mixed systems. These efforts helped identify conditions under which drug
co-hinding is likely to prevail in the environment

Experimental Methods

Materials and chemicals. Nalidixic acid (NA), Niflumic acid (NFA), Sulfamethoxazole (SMX), sodium chloride (NaCl), potassium hydroxide (KOH), sodium hydroxide (NaOH) and hydrochloric acid (HCl) were obtained from Sigma Aldrich, and were of analytical grade or better. The preparation and characteristics of goethite are detailed in the supporting information (SI).

Binding and co-binding experiments

Kinetic adsorption experiments were conducted in 125 mL Nalgene bottles
containing 0.5 g/L goethite in 10 mM NaCl under an atmosphere of $N_2(g)$. NA, NFA
and SMX concentrations were of 20 μM in both isolated (NA; NFA; SMX) and mixed
(NA+NFA; NA+SMX, NFA+SMX) systems. pH was adjusted using dilute NaOH or
HCl solutions to a pre-selected value. Aliquots were sampled during the course of the
experiments and filtered (0.2 μm) for analysis. Preliminary experiments showed that
adding the ligand simultaneously or sequentially after several hours of equilibration
had no significant effects on adsorption results.
Equilibrium adsorption experiments as a function of pH $(4 < pH < 9)$ were
conducted in 15 mL polypropylene tubes under an atmosphere of $N_2(g)$ to minimize
interferences with dissolved CO_2 at pH > 6.5 (Fig. S1). Adsorption isotherms were, in
turn, recorded at $pH = 6$ under $N_2(g)$ for (i) equimolar concentrations of NA and NFA
(0.1 - 40 μ M), (ii) [NA] _{tot} = 20 μ M and varying [NFA] _{tot} (0.1 - 40 μ M), and (iii)
[NFA] $_{tot}$ = 20 μM and varying [NA] $_{tot}$ (5 - 40 μM). The adsorbed amount was
calculated by depletion method. Desorption tests were also conducted at $pH = 11$ to
check the mass balance, and an average recovery of 99±2% for the investigated solutes
was obtained (see SI). Sorption and desorption experiments were performed at least
twice, and the reproducibility of the measurements was around 5% for NA and 10% for
NFA.
Aqueous concentrations of organic molecules were determined using a high
performance liquid chromatography (Waters 600 Controller) equipped with a

reversed-phase C18 column (250 mm×4.6 mm i.d., 5 μm) and a photodiode array
detector (Waters 996). The mobile phase was mixture of acetonitrile/water (60/40v/v)
contained 0.1% formic acid. The flow rate was set at 1 mL/min in isocratic mode. The
detector was set to 258 nm for NA, 283 nm for NFA and 270 nm SMX. All three
molecules could be analyzed with a single injection because they exhibited different
retention times (NA: 4.5 min; NFA: 10.1 min; SMX: 3.2 min).

ATR-FTIR spectroscopy and MCR analysis

Attenuated total reflectance-Fourier transform infrared (ATR-FTIR) spectra were recorded between in the 780-4000 cm⁻¹ region on an IS50 Nicolet spectrometer equipped with a KBr beam splitter and a liquid nitrogen cooled MCT detector. A nine-reflection diamond ATR accessory (Durasampl*IR* TM, Sens IR Technologies) was used for acquiring spectra of wet samples. The resolution of the single beam spectra was 4 cm⁻¹.

Sample preparation for the ATR-FTIR analysis was the same as for batch sorption experiments and has described in detail in our previous work²⁰. Spectra of goethite suspensions in 10 mM NaCl were also taken in the absence of NA and NFA and then subtracted from the spectra of sorbed NA and/or NFA in order to represent surface complexes only. Two series of experiments were conducted at pH=6 in 10 mM NaCl for 0.5 g/L goethite and (i) 0 < [NFA]_{tot} < 100 μ M with [NA]_{tot} = 100 μ M or (ii) 0 < [NA]_{tot} <100 μ M with [NFA]_{tot} = 100 μ M. Due to the relatively low solubility of both NA and NFA (see SI), 1 M NaOH was used to dissolve NA or NFA to ensure a high

concentration (10 mM) for ATR-FTIR analysis of NA and NFA aqueous solution. The
solid form of NA and NFA was also analysed using ATR-FTIR by loading powder on
the crystal, and then a drop of water was added to apply it more uniformly. Additionally,
the effect of pH (4-6) on NA and NFA sorption to goethite in 10 mM NaCl was
investigated for [NA] tot or [NFA] tot=100 μM as described in supporting information
(SI).
Selected sets of ATR-FTIR spectra in the 1200-1700 cm ⁻¹ region were then
analyzed by multivariate curve resolution (MCR) analysis ²¹ . These efforts extracted
spectral profiles and their relative concentrations (FTIR measurements cannot be used
to obtain absolute concentration values) of end-member components representing an
assemblage of the purest chemical species possible. Spectra sets were expressed in the
matrix A (m rows of wavenumber and n columns of measurements), and offset to zero
absorbance at 1700 cm ⁻¹ , where absorption by the wet mineral pastes is constant. The
spectra were expressed in terms of a linear combination of spectral profiles (ϵ) , akin to
molar absorption coefficients, and their concentration profiles (C), and are related by
$A=\epsilon X$ as in the Beer-Lambert law, such that that $\epsilon \ge 0$ and $C \ge 0$. Calculations of ϵ and
C were made with the MCR-ALS program ²¹ in the computational environment of
MATLAB (The Mathworks, Inc.). No assumptions regarding the spectroscopic
responses of the different species are made through this process.

DFT Calculations

DFT+D calculations were performed using an ab initio plane-wave pseudopotential
approach as implemented in VASP. ^{22,23} The Perdew-Burke-Ernzerhof (PBE)
functional ²⁴ was chosen to perform the periodic DFT calculations using the projector
augmented-wave method (PAW) ²⁵ and a cutoff of 400 eV. The dispersion forces were
taken into account using the Grimme D2 approach.26 To avoid the heavier
computational treatment of magnetic and electron-correlated iron oxides, we chose to
perform DFT calculations on two Al oxy-hydroxides (non-magnetic compounds): (i)
diaspore (α -AlOOH) which is the Al(III) isomorph of goethite, and (ii) gibbsite (AlOH ₂)
because the co-binding phenomenon is experimentally shown on this mineral surface
(See SI). This allowed also to perform more extensive calculations.
The bulk gibbsite and bulk diaspore were optimized and a (2x2) and a (4x4) cell was
chosen to build the basal surfaces, respectively. Then the molecules were optimized
separately in the same supercell as that used to model the surface, and the dimer was
also studied. Several protonation states of the NFA were considered. Since the
determination of adsorption free energy from water phase was not the aim of our study.
the solvent water molecules were not included in the calculations. The adsorption
energies computed here inform rather on the molecule-surface interaction strength. The
detailed calculation results are detailed in the SI.

Results and Discussion

Macroscopic assessment of NA and NFA binding. Binding kinetics of NA and NFA
in both single and binary systems followed pseudo-second-order kinetic model (Fig. S3)
and displayed comparable behaviours, with NA binding more strongly than NFA.
However, NFA loadings were considerably enhanced in the presence of NA, thus
providing a first line of evidence for synergetic intermolecular interactions at mineral
surfaces (SI). This can also be appreciated by ~4-fold slower adsorption rate of NFA in
the mixed system (pseudo-second order rate constant of 0.16 $\text{m}^2\text{/}\mu\text{mol}\cdot\text{min})$ than in the
isolated system (0.60 $\text{m}^2/\mu\text{mol}\cdot\text{min}).$ In contrast, mixed systems containing SMX did
not reveal any co-binding effects (Fig. S4).
NA and NFA binding at mineral surfaces in single system (Fig.1 for goethite)
follows the typical pH-dependent behaviour of carboxylic acids. 10,27-29 NA adsorption
was accordingly greatest under acid to circumneutral pH, where goethite surfaces are
positively charged, and NA carboxylate groups deprotonated (p K_a =6.19 for NA at
infinite dilution ^{30,} cf. Fig. S5). However, as NFA is a diprotic acid (pKa ₁ =2.28 and
$pKa_2=5.10$ at infinite dilution) ³¹ , it can exist as cationic, zwitterionic and anionic forms.
Only 23% of NFA was sorbed at acid pH and this percentage decreased with pH
increasing. Interestingly, binding of NA and NFA in mixed systems occurs over the
entire pH 4-9 range considered in this work (Figs. 1). This cooperative effect is more
pronounced for NFA because of its weaker adsorption in the isolated system (e.g.
increase of adsorption from 22% to 54% at pH 5). In addition, the pH-adsorption curve
of NFA (Fig.1b) becomes bell-shaped as in NA (Fig. 1a), suggesting that the NFA

binding to	goethite	surfaces	in the	binary	system	is	closely	related	to the	behavio	or of
NA bindin	g.										

Because this synergetic effect was observed for both molecules, two approaches
were adopted to study NFA and NA co-binding in mixed systems. Firstly, varying the
concentrations of NA and NFA, at ratio of 1:1 ([NFA] _{tot} = [NA] _{tot}), strongly points to
NA/NFA co-binding at goethite surfaces under a wide range of solute concentration
(0.1 to 40 μM), a range that notably partially overlaps with those in aquatic
environments (nM to several dozens of nM) ³⁻⁶ (Fig. 2a). Indeed, NFA and NA loadings
at pH 6 in isolated systems were lower than those measured in equimolar mixtures (Fig.
2a). Interestingly, by plotting the NFA loadings versus NA loadings in equimolar
mixtures, an excellent linear correlation was obtained (Fig. 2b, $[NFA]_{ads} = 0.6344$
$[NA]_{ads}$, $R^2 = 0.999$, fitted line was not shown). Secondly, varying $[NA]_{tot}$ at constant
[NFA] tot (20 μ M) and, conversely, varying [NFA] tot at constant [NA] tot (20 μ M) showed
that increasing surface loadings of one ligand increases the other. However, a plateau
was reached for [NFA] _{ads} where [NA] _{tot} varies, which is likely to have arisen from
molecular layers acting as steric or electrostatic barriers preventing additional
binding. ³²

Molecular investigations of co-binding. Vibration spectroscopy and density functional theory (DFT) calculations were used to provide clues on the mechanisms through which NA and NFA bind and co-bind at goethite surfaces. We note that DFT calculations were performed on diaspore (Fig. 3), which is the Al(III) isomorph of

goethite, to avoid the otherwise heavier computational treatment of magnetic and electron-correlated iron oxides. The (110) face was chosen to emulate the dominant crystallographic face of the goethite particles under study.

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The fingerprint region of these molecules (1200-1800 cm⁻¹; Fig. 4, see also band assignments in SI) showed a 25 cm⁻¹ blue shifts in C-O stretching modes (v_{COO}), while no obvious shift for the ring modes v_{ring} was observed. This suggests direct interactions of carboxyl groups with goethite but little interaction with the aromatic and pyridine rings during the sorption of NA and NFA in single system^{13,33}. Though vibration spectroscopy suggests both metal- and hydrogen-bonding for NA, DFT calculations suggest that hydrogen bonding is the preferred binding mode for NA (-0.34 eV vs. +0.44 eV for inner sphere complexation) and that it is 0.37 eV more favourable than NFA (+0.03 eV). Thus while both complexes are stabilised by direct hydrogen bonds between carboxyl groups and surface hydroxo groups, NA binding is made stronger by a vicinal carbonyl of the pyridine ring and involves a hydrogen-bond cycle between the molecule and two surface water molecules (Fig.3). In contrast, this cycle is not only absent in NFA but when we forcefully hydrogen bonded NFA with an adsorbed water simulations showed that this water reoriented itself towards a neighboring water molecule. The weak nature of NFA binding can even be compared to those of monocarboxylic acids (e.g. acetate or benzoate¹³).

238	Vibration spectra of mixed NA+NFA systems exposed to goethite (Fig. 4) showed
239	that increasing NFA concentrations (0, 10, 20, 50 and 100 $\mu M)$ with [NA] $_{tot}$ = 100 μM
240	systematically increased the intensities of the characteristic bands of NFA
241	$(v_{COO,as}=1480-1560 \text{ cm}^{-1})$, yet the resulting spectra cannot be represented as simple
242	linear combinations of the isolated goethite-NA and goethite-NFA systems (Fig. S7a).
243	For instance, the ring mode ($v_{C=C,ring}$) of NA was shifted from 1578 cm ⁻¹ to 1522 cm ⁻¹
244	and that of NFA was split into two bands (1335 and 1348 cm ⁻¹) suggesting perturbation
245	of C-C stretches and/or C-H bends of the aromatic and pyridine rings ^{34,35} , and thus
246	formation of dimer involving the aromatic and pyridine rings of NFA and NA. These
247	observations also hold for the converse experiments where NA concentrations (0, 10,
248	20, 50 and 100 μ M) are increased with [NA] _{tot} = 100 μ M (Fig. S7b).
249	A multivariate curve resolution (MCR) analysis ²¹ of these spectral sets provided
250	further insight into the nature of NFA and NA co-binding. MCR decomposed each
251	spectral sets into two separate spectral components (Figs. 5 a,b) representing the purest
252	extractable mineral-bound NFA and NA complexes (MCR I) and those under
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253	competing systems (MCR II). The related concentration profiles (Fig. 5c) revealed that
253254	competing systems (MCR II). The related concentration profiles (Fig. 5c) revealed that addition of NA to mineral-bound NFA was more effective at altering the spectral
254	addition of NA to mineral-bound NFA was more effective at altering the spectral
254 255	addition of NA to mineral-bound NFA was more effective at altering the spectral profile of NFA than the converse addition of NFA to mineral-bound NA. Still, as the
254255256	addition of NA to mineral-bound NFA was more effective at altering the spectral profile of NFA than the converse addition of NFA to mineral-bound NA. Still, as the resulting MCR II components are markedly similar, our results suggest that the

In line with the concept that NA enhances NFA binding, DFT calculations reveal that NFA binding to a hydrogen-bound NA on diaspore is energetically favourable (-0.21 eV). The resulting dimer formed via favourable hydrogen bonding and van der Waals interactions by -0.50 eV, and the two COOH moieties of this dimer are parallel with one another thus increasing the strengths of its interactions with mineral surfaces (*cf.* SI for more information and Figures S10-S15). As such, recalling that NA binding is favourable by -0.34 eV, binding of NFA to a pre-sorbed NA should be favourable by -0.55 eV. In comparison, NA and NFA binding at different locations on the same diaspore surface are favourable by only -0.30 eV and the formation an unbound NA/NFA dimer is favourable by -0.50 eV.

Because the main mode of attachment for NA is achieved via hydrogen bonding at circumneutral conditions, this phenomenon is not only limited to strongly reactive faces of minerals, such as the (110) face of goethite/diaspore or edges of clays, but also on the basal planes of minerals. The planes are of widespread occurrence in platy metal (oxy)(hydr)oxides as well as phyllosilicates (*e.g.* clays) and typically display (hydr)oxo groups that are strongly resilient to ligand exchange, yet are active hydrogen bonding sites. To illustrate this point further the SI contains further details on the energetics of NA/NFA co-binding on the basal plane of gibbsite, an important aluminium hydroxide in natural but also industrial settings. Gibbsite gives more weight to our demonstration, since the NA/NFA co-binding is experimentally shown to occur on this mineral surface (See Fig. S16).

Those results also fall precisely in line with those obtained for diaspore, and suggests the possibility in generalizing our finding to an even broader range of minerals and particles which capable of stabilizing NA-like molecules via hydrogen bonding. Our calculations consequently lend strong independent support for the concept that NA and NFA co-bind at mineral surfaces of even contrasting structure, and that a dimer-type species stabilized by intermolecular hydrogen bonding and van der Waals interactions could be responsible for this phenomenon.

Implications for transport of pharmaceutical compounds in nature. Our concerted macroscopic and molecular efforts both provide evidence that NFA-NA interactions mutually enhance binding at mineral surfaces such as goethite. This cooperative effect is more pronounced for NFA because of its intrinsically weaker affinity for mineral surfaces, and occurs under environmentally relevant conditions of drug concentration and pH. Vibration spectroscopic data show that addition of NA effectively alters the nature of mineral-NFA binding but that converse addition of NFA to mineral-bound NA results in a less dramatic change in the nature of NA binding. In support to these finding DFT calculations showed that NFA binding on mineral faces of even strongly contrasting structures is thermodynamically favoured only when NA is pre-adsorbed either metal-bonded or hydrogen-bonded. This favoured form of binding could be explained by the formation of a NFA-NA dimer stabilised by hydrogen bonding and van der Waals interactions. The lack of co-binding seen in SMX also suggests the

importance of understanding drug interactions in aqueous solutions, a finding that also calls for new studies along these lines.

This study is the first to show that mineral-bound antibiotic molecules can be specific adsorption sites for other antibiotic molecules, and that layered-like coatings involving anti-inflammatory agents may even form at mineral surfaces. As water resources are exposed to complex mixtures of chemicals³⁶, including natural organic matter and metal ions, additional efforts resolving the underlying principles governing cooperative sorption should be made to accurately assess the fate of co-existing contaminants in the environment. This becomes even more so urgent under the growing number aquatic ecosystems and groundwater systems exposed to emerging contaminants including non-prescription drugs, antibiotics, hormones and prescription drugs^{36,37}. As a result, recognising the importance of drug co-binding at mineral surfaces, and the conditions where it does not occur such as in the case of SMX, is key to the successful development of models for predicting the fate of these contaminants, and for guiding policies on actions needed to mitigate this growing environmental problem.

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325	Supporting Information Available
326	Details of the materials used in this study, mineral characterization, analytical and DFT
327	methods, and additional data are available in the Supporting Information. This
328	information is available free of charge via the Internet at http://pubs.acs.org/ .
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330 References

- 331 (1) Lapworth, D. J.; Baran, N.; Stuart, M. E.; Ward, R. S. Emerging organic
- contaminants in groundwater: a review of sources, fate and occurrence. *Environ*.
- *Pollut.* **2012**, *163*, 287–303.
- 334 (2) Kolpin, D. W.; Furlong, E. T.; Meyer, M. T.; Thurman, E. M.; Zaugg, S. D.;
- Barber, L. B.; Buxton, H. T. Pharmaceuticals, hormones, and other organic
- wastewater contaminants in US streams, 1999-2000: A national reconnaissance.
- 337 Environ. Sci. Technol. **2002**, 36 (6), 1202–1211.
- 338 (3) Hernando, M. D.; Mezcua, M.; Fernández-Alba, A. R.; Barceló, D.
- Environmental risk assessment of pharmaceutical residues in wastewater
- effluents, surface waters and sediments. *Talanta* **2006**, *69* (2), 334–342.
- 341 (4) Heberer, T. Occurrence, fate, and removal of pharmaceutical residues in the
- aquatic environment: a review of recent research data. *Toxicol. Lett.* **2002**, *131*
- 343 (1), 5–17.
- 344 (5) Fatta-Kassinos, D.; Meric, S.; Nikolaou, A. Pharmaceutical residues in
- environmental waters and wastewater: Current state of knowledge and future
- research. Anal. Bioanal. Chem. **2011**, 399 (1), 251–275.
- 347 (6) Gothwal, R.; Shashidhar, T. Antibiotic Pollution in the Environment: A Review.
- 348 *CLEAN--Soil, Air, Water* **2015**, *43* (4), 479–489.
- 349 (7) Gu, C.; Karthikeyan, K. G. Sorption of the antimicrobial ciprofloxacin to
- aluminum and iron hydrous oxides. Environ. Sci. Technol. 2005, 39 (23),
- 351 9166–9173.

- Wang, Y.; Newman, D. K. Redox reactions of phenazine antibiotics with ferric
- 353 (hydr) oxides and molecular oxygen. Environ. Sci. Technol. 2008, 42 (7),
- 354 2380–2386.
- 355 (9) Xu, X.-R.; Li, X.-Y. Sorption and desorption of antibiotic tetracycline on marine
- sediments. *Chemosphere* **2010**, 78 (4), 430–436.
- 357 (10) Paul, T.; Liu, J.; Machesky, M. L.; Strathmann, T. J. Adsorption of zwitterionic
- fluoroquinolone antibacterials to goethite: A charge distribution-multisite
- complexation model. *J. Colloid Interface Sci.* **2014**, 428, 63–72.
- 360 (11) Pouliquen, H.; Le Bris, H. Sorption of oxolinic acid and oxytetracycline to
- marine sediments. *Chemosphere* **1996**, *33* (5), 801–815.
- 362 (12) Kulshrestha, P.; Giese, R. F.; Aga, D. S. Investigating the molecular interactions
- of oxytetracycline in clay and organic matter: insights on factors affecting its
- mobility in soil. *Environ. Sci. Technol.* **2004**, *38* (15), 4097–4105.
- 365 (13) Norén, K.; Persson, P. Adsorption of monocarboxylates at the water/goethite
- interface: The importance of hydrogen bonding. Geochim. Cosmochim. Acta
- **2007**, *71* (23), 5717–5730.
- 368 (14) Sun, D.; Zhang, X.; Wu, Y.; Liu, T. Kinetic mechanism of competitive
- adsorption of disperse dye and anionic dye on fly ash. Int. J. Environ. Sci.
- 370 *Technol.* **2013**, 10 (4), 799–808.
- 371 (15) Conkle, J. L.; Lattao, C.; White, J. R.; Cook, R. L. Competitive sorption and
- desorption behavior for three fluoroquinolone antibiotics in a wastewater
- treatment wetland soil. *Chemosphere* **2010**, *80* (11), 1353–1359.

- 374 (16) Xing, B.; Pignatello, J. J. Competitive sorption between 1, 3-dichlorobenzene or
- 2, 4-dichlorophenol and natural aromatic acids in soil organic matter. *Environ*.
- 376 Sci. Technol. 1998, 32 (5), 614–619.
- 377 (17) Jones-Lepp, T. L.; Alvarez, D. A.; Englert, B.; Batt, A. L. Pharmaceuticals and
- Hormones in the Environment. *Encycl. Anal. Chem.* **2009**.
- 379 (18) Zhang, T.; Li, B. Occurrence, Transformation, and Fate of Antibiotics in
- Municipal Wastewater Treatment Plants. Crit. Rev. Environ. Sci. Technol. 2011,
- 381 *41* (11), 951–998.
- 382 (19) Jagtap, S.; Yenkie, M. K.; Labhsetwar, N.; Rayalu, S. Fluoride in drinking water
- and defluoridation of water. *Chem. Rev.* **2012**, *112* (4), 2454–2466.
- 384 (20) Marsac, R.; Martin, S.; Boily, J.-F.; Hanna, K. Oxolinic acid binding at goethite
- and akaganéite surfaces: implications for aquaculture-induced pollution.
- 386 Environ. Sci. Technol. **2016**, 50 (2), 660–668.
- 387 (21) Jaumot, J.; Gargallo, R.; de Juan, A.; Tauler, R. A graphical user-friendly
- interface for MCR-ALS: a new tool for multivariate curve resolution in
- 389 MATLAB. Chemom. Intell. Lab. Syst. **2005**, 76 (1), 101–110.
- 390 (22) Kresse, G.; Hafner, J. Ab initio molecular dynamics for liquid metals. *Phys. Rev.*
- 391 *B* **1993**, *47* (1), 558–561.
- 392 (23) Kresse, G.; Hafner, J. Norm-conserving and ultrasoft pseudopotentials for
- first-row and transition elements. J. Phys. Condens. Matter 1994, 6 (40), 8245.
- 394 (24) Perdew, J. P.; Burke, K.; Ernzerhof, M. Generalized gradient approximation
- 395 made simple. *Phys. Rev. Lett.* **1996**, 77 (18), 3865–3868.

- 396 (25) Blöchl, P. E. Projector augmented-wave method. Phys. Rev. B 1994, 50 (24),
- 397 17953–17979.
- 398 (26) Grimme, S.; Antony, J.; Ehrlich, S.; Krieg, H. A consistent and accurate ab initio
- parametrization of density functional dispersion correction (DFT-D) for the 94
- 400 elements H-Pu. J. Chem. Phys. **2010**, 132 (15), 154104.
- 401 (27) Rusch, B.; Hanna, K.; Humbert, B. Sorption and transport of salicylate in a
- porous heterogeneous medium of silica quartz and goethite. Environ. Sci.
- 403 *Technol.* **2010**, *44* (7), 2447–2453.
- 404 (28) Paul, T.; Machesky, M. L.; Strathmann, T. J. Surface complexation of the
- zwitterionic fluoroquinolone antibiotic ofloxacin to nano-anatase TiO2
- 406 photocatalyst surfaces. *Environ. Sci. Technol.* **2012**, *46* (21), 11896–11904.
- 407 (29) Pei, Z.; Shan, X. Q.; Kong, J.; Wen, B.; Owens, G. Coadsorption of
- 408 ciprofloxacin and Cu(II) on montmorillonite and kaolinite as affected by
- solution pH. *Environ. Sci. Technol.* **2010**, *44* (3), 915–920.
- 410 (30) Ross, D. L.; Riley, C. M. Aqueous solubilities of some variously substituted
- 411 quinolone antimicrobials. *Int. J. Pharm.* **1990**, *63* (3), 237–250.
- 412 (31) Takács-Novák, K.; Tam, K. Y. Multiwavelength spectrophotometric
- 413 determination of acid dissociation constants: Part V: microconstants and
- tautomeric ratios of diprotic amphoteric drugs. J. Pharm. Biomed. Anal. 2000,
- *21* (6), 1171–1182.
- 416 (32) Liu, S. Cooperative adsorption on solid surfaces. J. Colloid Interface Sci. 2015,
- *450*, 224–238.

418	(33)	Madey, T. E.; Yates Jr, J. T. Vibrational spectroscopy of molecules on surfaces
419		Springer Science & Business Media, 2013; Vol. 1.
420	(34)	Akyuz, S.; Akyuz, T. FT-IR spectroscopic investigations of adsorption of 2-
421		3-and 4-pyridinecarboxamide on montmorillonite and saponite from Anatolia
422		Vib. Spectrosc. 2006 , 42 (2), 387–391.
423	(35)	Balci, K.; Akkaya, Y.; Akyuz, S. An experimental and theoretical vibrational
424		spectroscopic study on niflumic acid, a non-steroidal anti-inflammatory drug
425		Vib. Spectrosc. 2010 , 53 (2), 239–247.
426	(36)	Schwarzenbach, R. P.; Escher, B. I.; Fenner, K.; Hofstetter, T. B.; Johnson, C. A.;
427		von Gunten, U.; Wehrli, B. The challenge of micropollutants in aquatic systems
428		Science 2006 , 313 (5790), 1072–1077.
429	(37)	Rotter, S.; Gunold, R.; Mothes, S.; Paschke, A.; Brack, W.; Altenburger, R.
430		Schmitt-Jansen, M. Pollution-Induced Community Tolerance To Diagnose
431		Hazardous Chemicals in Multiple Contaminated Aquatic Systems. Environ. Sci.
432		Technol. 2015, 49 (16), 10048–10056.
433		
434		

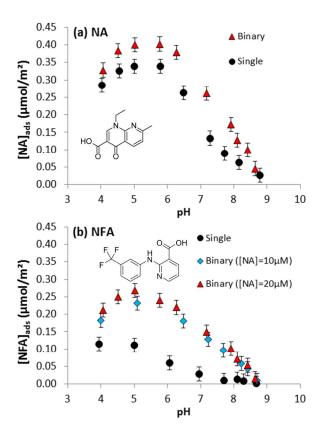
Figure captions

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- Figure 1. Left: pH-adsorption edges of a) NA single ([NA]_{tot} = 20 μ M) and binary
- 438 ([NA]_{tot} = [NFA]_{tot} = 20 μ M) and b) NFA single ([NFA]_{tot} = 20 μ M) and binary
- ([NFA]_{tot} = $20 \mu M$; [NA]_{tot} = $10 \text{ and } 20 \mu M$) systems on goethite, with 10 mM NaCl.
- Right: molecular structures of NA and NFA.
- Figure 2. a) NA and NFA sorption to goethite for single systems (full symbols) and
- binary system where $[NA]_{tot} = [NFA]_{tot}$ (empty symbols). NA and NFA concentrations
- were varied from 0.1 to 40 μM. For the sake of readability, the behaviour at very low
- concentrations was shown in the insert. b) [NFA]_{ads} vs [NA]_{ads} at three experimental
- conditions: (i) varying both compounds from 0 to 40 μ M (black), (ii)[NFA]_{tot} = 20 μ M,
- 446 $0 < [NA]_{tot} < 40 \mu M \text{ (red)}, \text{ and (iii) } [NA]_{tot} = 20 \mu M, 0 < [NFA]_{tot} < 40 \mu M \text{ (blue)}.$
- Figure 3. NA and NFA molecules co-adsorbed on the diaspore surface, with NA
- adsorbed as (*left*) inner sphere (E_{ads} (NFA/NA) = 0.07 eV, athermic process) and (*right*)
- outer sphere (E_{ads} (NFA/NA) = -0.21 eV). A negative energy indicates an exothermic
- 450 process.
- Figure 4. ATR-FTIR spectroscopy on goethite. (a) from top to bottom: NA single
- system, dissolved NA (NA_(aq) in 1 M NaOH), NA-NFA binary system ([NA]_{tot} = 100
- 453 μ M, $10 < [NFA]_{tot} < 100 \mu$ M; the arrows show increasing [NFA]_tot), NFA single system,
- dissolved NFA (NFA_(aq) in 1 M NaOH). Bold and thin dashed lines show characteristic
- bands of NA_(aq) and NFA_(aq), respectively. Numbers denote [NFA]_{tot}. Spectra were
- normalized according to the band at 1448 cm⁻¹, since NA is the major component.

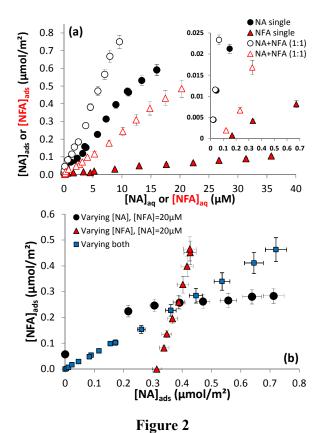
457	Figure 5. MCR-extracted spectral profiles from FTIR spectra of (a)100 μ m NFA + NA ,
458	(b) $100~\mu m$ NA + NFA, both including reference spectra, and (c) associated
459	concentration profiles corresponding to components MCR I and II. These concentration
460	profiles underscore the larger propensity of NA at displacing bound NFA, than NFA at
461	displacing NA.
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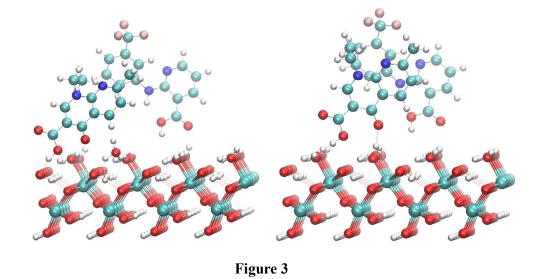


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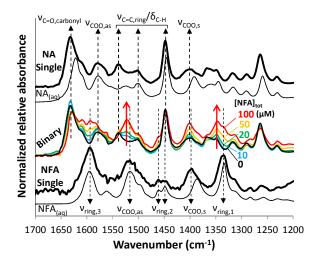
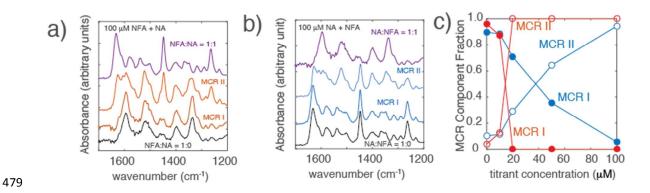


Figure 4

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480 Figure 5

